

Amazon forest biomass: intra- and interspecific variability in wood density drive divergences in Brazil's far north

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Wood density (WD) is an important functional trait of tree species. Understanding spatial WD variability as a function of environmental determinants improves our ability to estimate carbon stocks in the woody biomass of tropical forests. However, the role of each environmental variable affecting the intra- and interspecific variability of WD is not entirely clear for most forest ecosystems. In Amazonia there are recurrent uncertainties in estimates of regional woody biomass. The aim of the study was to investigate the effects of environmental conditions on the intra- and interspecific variability of WD for tree assemblages in forests of the northern Brazilian Amazon. A single sample was extracted from each of 680 individuals (108 species, 82 genera, 38 families; stem diameter ≥ 10 cm) dispersed among 129 plots distributed along a hydro-edaphic gradient. General community-averaged WD (0.703 ± 0.133 g cm⁻³; range: 0.203 to 1.102 g cm⁻³) was high in relation to other Amazonian areas because 62% of the species and 69% of the sampled individuals had high WD values (>0.650 g cm⁻³). Altitude (a proxy for drainage), clay and soil micronutrient content explained 23% of the spatial variation in WD. Partitioning WD variation into species-substitution (turnover) and intraspecific-variation components slightly increased the explanatory power to 26%. The analysis of interspecific variability showed that forests occurring in seasonally flooded areas are characterized by tree assemblages with species tolerant to P-poor soils, where mean WD (0.742 g cm⁻³) is about 4% higher than the mean (0.713 g cm⁻³) for tree assemblages on unflooded uplands where soils have less limitations from nutrient poverty. Our results represent an improvement in the estimates of biomass because they promote adjustments (1.4%-16.3%) to the previous estimates of woody biomass in the northern Brazilian Amazon forests considering different environmental conditions.

Keywords: Basic Density, Maracá, Roraima, Seasonal Forests, Wood Specific Gravity

Introduction

Wood density, also termed specific gravity, is a functional trait strongly associated with interspecific variation in tree growth, architecture, mortality and resistance to

water stress (Sarmiento et al. 2011, Oliveira et al. 2019). Basic wood density, here abbreviated as “WD”, refers to oven-dry weight divided by saturated volume, this being the most appropriate density mea-

sure for calculating the biomass of trees from estimates of live volume. WD is one of the main variables used in allometric models for indirect calculations of woody biomass to estimate tropical forest carbon stocks (Chave et al. 2014, Dias & Marengo 2016).

While an accurate knowledge of WD variability is recognized as a functional parameter that significantly improves tropical-forest biomass prediction, it is also considered to be one of the main sources of error in the estimates (Baker et al. 2004, Chave et al. 2006). The main causes for WD to be a source of error propagation in tropical-forest biomass estimation are associated with (i) difficulties in collecting a sufficiently large number of samples to represent the main species (Fearnside 1997), (ii) obtaining samples that represent the intra- and inter-intraspecific spatial variability between species and different forest ecosystems (Muller-Landau 2004), and (iii) standardization problems in sample collection and methodological analysis (Williamson & Wiemann 2010). The difficulty of measuring WD leads most investigations involving tree biomass estimation in tropical forests

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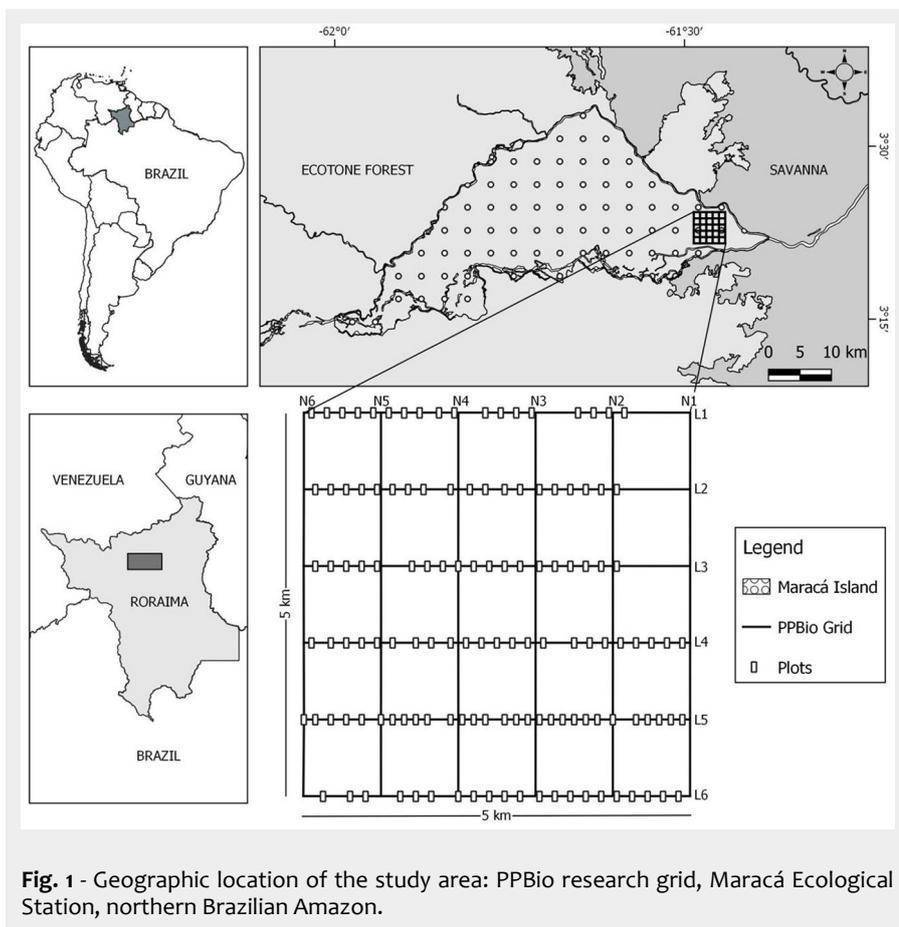


Fig. 1 - Geographic location of the study area: PPBio research grid, Maracá Ecological Station, northern Brazilian Amazon.

highly dependent on environmental conditions (Baker et al. 2004, Cosme et al. 2017). However, for ecotone forests in northern Brazilian Amazonia the effect of the hydro-edaphic gradient on WD variation has not yet been assessed, and there is a paucity of information on the biometric and functional variables needed to estimate regional biomass and carbon stocks in these forests (Barbosa et al. 2019). Accounting for the different underlying variables driving WD variation is therefore important for understanding and accurate models to estimate biomass and carbon stocks in the ecotone forests in the northern Brazilian Amazonia.

The aim of the present study was to investigate community average WD and its determinants in the tree assemblages of ecotone forests in northern Brazilian Amazon. The specific questions for our investigation were: (i) how does the hydro-edaphic gradient drive the spatial variability of wood density in northern Amazonian forests? and (ii) how do the relative contributions of intra- and interspecific WD variation respond to the hydro-edaphic gradient? The development of a regional database for the wood densities of tree species in northern Amazonian forests will provide a better way to calculate woody biomass and carbon stocks in this poorly studied ecological region where current estimates are based on databases derived from regions where climatic, hydro-edaphic and vegetation conditions are different.

Material and methods

Study area

The study was carried out in the PPBio (Biodiversity Research Program) research grid installed in the eastern portion of the Maracá Ecological Station (<https://ppbio.inpa.gov.br/sitios/maraca>), a Brazilian protected area located in the northern portion of the state of Roraima ($03^{\circ}15' - 03^{\circ}35' N$ and $61^{\circ}22' - 61^{\circ}58' W$), about 135 km from Boa Vista, the state capital (Fig. 1). The ecological station is formed by Maracá Island and other fluvial islets (hereafter “Maracá”); it is 60 km long and 15-25 km wide and has a total area of ~101,000 ha (Silva et al. 2019). Maracá is located in an area of climatic transition between the Köppen-classification subtypes Aw and Am, where rainfall and average annual temperature are 2086 mm and 26 °C, respectively (Couto-Santos et al. 2014). The rainy season (> 300 mm month⁻¹) occurs mainly between May and August, while the driest period is between December and March (< 100 mm month⁻¹ – Carvalho et al. 2018).

The eastern portion of Maracá is an ecotone zone characterized by a savanna-forest contact zone where the forest types are determined by different hydro-edaphic conditions (Nascimento et al. 2017, Villacorta et al. 2022). The forest mosaic as a whole is an accurate representation of the large ecotone forest area in the Northern Brazil-

to use WD values extracted from global repositories (Zanne et al. 2009). Although biomass estimates carried out using the WD values from global databases can be considered an advance, this practice can lead to divergences in estimates of carbon stocks (Mitchard et al. 2014). Overestimation of WD by up to 16% can occur at the level of a species community (Ramanananthoandro et al. 2015).

Amazonia is the world’s most diverse and largest continuous area of tropical forest (Esquivel-Muelbert et al. 2019). Despite recent advances in carbon and biomass stocks estimation, there are still great uncertainties associated with calculations of biomass loss and accumulation in Amazonia’s different forest ecosystems (Tejada et al. 2019). Divergences in the regional spatial patterns of forest dynamics in Amazonia have been attributed to WD variability, where low-WD forests in northwestern Amazonia have high mortality and fast turnover, while northeastern Amazonia has high-WD forests with low mortality (Chao et al. 2008). There is a consensus that environmental variables are the main factors affecting the inter- and intra-specific spatial patterns in WD (Siliprandi et al. 2016, Poorter et al. 2019). Large environmental variability (e.g., soil type, climatic seasonality, water table deep and flooding periodicity) in Amazonia impacts functional traits of tree species at various levels (Quezada et al. 2012, Esteban et al. 2021). Consequently, environmental heterogeneity gen-

erates variations in structural and functional traits, introducing inaccuracies into general allometric models intended for estimating carbon stocks based on forest biomass (Chave et al. 2014).

Some studies in Amazonia have shown that intraspecific variations can reflect phenotypic plasticity of the species caused by edaphic factors (Parolin & Worbes 2000), which affect changes in structure and tree species composition (Muller-Landau 2004, Costa et al. 2023), where WD values are inversely related to soil fertility (Baker et al. 2004). Flooding periodicity (intermittent drainage) also can affect the wood density, because habitat specific WD is the key to quantifying above-ground woody biomass and carbon differences between flooded and non-flooded forests due to site specificities, such as taxonomic groups and tree dimensions (Bredin et al. 2020).

Recent studies have identified a hydro-edaphic gradient acting as an environmental filter that selects tree species, and shapes the ecotone forest’s structure in northern Amazonia (Nascimento et al. 2017, Silva et al. 2021). These ecotone forests occur where there is high environmental heterogeneity that results in a large mosaic of different forest types (ombrophilous, semideciduous and deciduous) (Villacorta et al. 2022). This results in forests with structure and floristic composition that are distinct from other Amazonian forests. Intra- and interspecific variability in WD is higher in forests that are

ian Amazonia (Barbosa et al. 2019). The forest types are defined by the Brazilian Vegetation Classification System as ombrophilous and seasonal (semideciduous and deciduous) forests (Brazil-IBGE 2012). The forest formations occur on a variety of different reliefs and soils ranging from Typic Tropaquept (lowland; hydromorphic; seasonally flooded) to Oxic Haplustult and Typic Haplustalf, both well drained (Nortcliff & Robison 1998).

Sampling design

Between December 2015 and January 2016, 129 plots (50 × 10 m, 6.45 ha in total) were installed along the six east-west trails of the PPBio grid (Fig. 1). The distance between individual plots was 150 m and was measured using PPBio distance pickets established every 50 m along the trails that cross the PPBio grid; all pickets are georeferenced in UTM coordinates, including altitude (m a.s.l.) as a topographic variable (<https://ppbiodata.inpa.gov.br/metacat/metacat/menger.192.1/default>). Aquatic environments (marshes) and savannas were excluded from the sampling because they are not forest ecosystems.

All trees with a stem diameter ≥10 cm in each plot were inventoried and marked with numbered aluminum tags. POM (point of measure) height was adopted as a reference for measuring stem diameter. Most diameters were measured at 1.30 m above the ground, the exceptions being when the tree individual had buttress roots or other problems (e.g., bifurcated trees), making it necessary to reconfigure the POM to 0.5 m above the physical impediment, according to the tree-measurement protocol adopted by the study (https://ppbio.inpa.gov.br/sites/default/files/Protocolo_estrutura_vegetacao_2014_o.pdf). A diameter tape (model 283D/5m) was used to measure stem diameters. Plots have been censused annually since 2015, and all biometric measures are available through the Mendeley Data repository (<https://data.mendeley.com/datasets/8cdwkhcsy7/2>) and the ForestPlots platform (<https://www.forestplots.net/>) under the codes ETA, ETB, ETC, ETD, ETE and ETF.

All individuals were morphotyped, and botanical material was collected to enable the taxonomic identification of all trees to the lowest possible taxonomic level. Specimens were prepared and deposited in the INPA, MIRR and UFRR Herbaria (herbarium acronyms follow Thiers 2020). Scientific names were verified and corrected through the Brazilian Flora Species list. Family-level circumscriptions followed APG-IV (2016). The species list for each plot (Silva et al. 2019) can be freely accessed at the Global Biodiversity Information Facility - GBIF (<https://doi.org/10.15468/xa5lrb>).

Wood density

Fieldwork was carried out in all 129 plots in two stages: January 2018 (269 samples in the 3rd tree census) and January 2019 (411

samples in the 4th tree census). Both were carried out during the dry season in order to maintain the same weather-pattern data collection. The collection of 680 samples (108 species, 82 genera, 38 families; except palms) was determined randomly considering 25% of the individual trees in each plot, regardless of the diameter class or species. A single sample was removed per stem (at ~1.30 m above the ground) using an increment borer (Haglof Borer Auger – 400 mm in length and 5.15 mm in diameter), as specified in the RAINFOR field manual to measure wood density in tropical forest trees ([http://www.rainfor.org/upload/ManualsEnglish/wood_density_english\[1\].pdf](http://www.rainfor.org/upload/ManualsEnglish/wood_density_english[1].pdf)).

Collected samples were packed in plastic bags and kept in a cooler with ice to avoid dehydration. All samples were sent to the laboratory and separated into two segments: (i) core wood (material corresponding to heartwood + sapwood); and (ii) bark (internal + external = material corresponding to the space between the rhytidome and the cambium). The thickness (mm) and the wood densities (g cm⁻³) of the two segments were measured separately to obtain individual results and normalize the wood density values for each sampled tree applying a weighted average between bark and core wood. We followed this methodological step because the bark is generally an omitted component in wood-density studies (Williamson & Wiemann 2010). However, the bark is an important component in Amazon forest trees (Staver et al. 2020), and it should not be neglected in studies of wood density (Nogueira et al. 2007). The weighted average of the wood densities of the two components (using the radius = thickness of the bark and + core wood length as a single linear representation of the stem cross-sectional) is the most suitable value for use in estimates of woody biomass and carbon. This methodological approach was conceived with the aim of reducing the error associated with the wood-density values used to calculate the biomass and carbon of individuals dispersed in tropical ecosystems with high environmental variability. The weighted average WD values were calculated using the equation below (eqn. 1):

$$WD_w = \frac{(Bark_{th} \cdot WD_{bark}) + (Core_{th} \cdot WD_{core})}{R_{tree}} \quad (1)$$

where WD_w is the weighted average of the wood density (g cm⁻³), $Bark_{th}$ is the thickness of the bark (cm), WD_{bark} is the bark wood density (g cm⁻³), $Core_{th}$ is the wood core length (cm), WD_{core} is the core wood density (g cm⁻³), and R_{tree} is the tree radius estimated as thickness of the bark (cm) + wood core length (cm).

To calculate wood density (bark and core wood), the ratio of oven-dried mass (g) divided by the saturated (green) volume (g cm⁻³) was used (Williamson & Wiemann 2010). Saturated volume of each sampled segment was estimated using a graduated cylinder with distilled water on a precision

balance (0.001 g). Each sampled segment was immersed in water, where its weight was measured by the displaced volume, considering the water density equal to 1 g cm⁻³, following the methodology indicated by Williamson & Wiemann (2010). All sampled segments were dried in an oven (100 ± 2 °C) until constant weight was reached. As suggested by Barbosa et al. (2017), we associated all wood density values using the categories adopted by the Brazilian Institute of the Environment and Renewable Natural Resources (IBAMA–Sternadt 2001): very-soft wood species (< 0.350 g cm⁻³), soft (0.351–0.500 g cm⁻³), medium-soft (0.501–0.650 g cm⁻³), medium-hard (0.651–0.800 g cm⁻³), hard (0.801–0.950 g cm⁻³) and very-hard (> 0.951 g cm⁻³). This categorical distribution is useful for understanding the tendency of different sites to host species and/or individuals with higher (> 0.650 g cm⁻³) or lower (≤ 0.650 g cm⁻³) wood density. The raw dataset obtained in the current study can be freely assessed in international repositories (Farias et al. 2020).

Environmental variables

Soil samples for each plot were obtained by collection of two superficial subsamples (0–20 cm deep). The two sub-samples were homogenized (~500 g), air-dried and sieved (2 mm mesh). Values of edaphic variables were determined: clay content (%), organic matter content (mg kg⁻¹), phosphorus content (mg kg⁻¹), sum of exchangeable bases (K⁺ + Ca⁺⁺ + Mg⁺⁺ – cmol kg⁻¹), sum of micronutrients (Fe, Zn, Mn, Cu and B – mg kg⁻¹), and pH (H₂O). Analyses were carried out following the analytical methods in the Brazilian Agricultural Research Corporation soil chemistry analysis manual (*Manual de Métodos de Análise de Solo* – <http://ainfo.cnptia.embrapa.br/digital/bitstream/item/104933/1/Manual-de-Mtodos-de-Anilise-de-Solo.pdf>). A dataset for chemical and physical soil analyses is freely accessible via Mendeley Data (<https://data.mendeley.com/datasets/gfw5ccbrsz/2>). Altitude (m a.s.l., used here as a proxy for drainage) and planimetric coordinates (UTM) were obtained from the PPBio data repository (<https://ppbiodata.inpa.gov.br/metacat/metacat/menger.192.1/default>). Plots at ≤ 65 m a.s.l. were considered to be poorly drained (seasonally flooded), while those at > 65 m a.s.l. were considered to be free of flooding, following the criteria adopted by Villacorta et al. (2022) for the eastern portion of Maracá. These categories correspond, respectively, to forests on shallow (≤ 65 m a.s.l.) and deep (> 65 m a.s.l.) water tables following the topographic classification adopted by Esteban et al. (2021).

Statistical analysis

Regression models were used to investigate the relationship between the composition of the tree community in terms of wood density and considered environmental variables. Community composition was quantified using the specific average of

Tab. 1 - Wood-density values distributed among the Brazilian Institute of the Environment and Renewable Natural Resources (IBAMA) categories: very-soft wood species (< 0.350 g cm⁻³), soft (0.351-0.500 g cm⁻³), medium-soft (0.501-0.650 g cm⁻³), medium-hard (0.651-0.800 g cm⁻³), hard (0.801-0.950 g cm⁻³) and very-hard (> 0.951 g cm⁻³). (Field samples): number of individuals sampled in each category; (Total trees): total trees present in the forest survey (except palms and indeterminate individuals); (Bark WD): bark density; (Core WD): density of sapwood + heartwood; (WD): weighted average between Bark WD and Core WD.

WD Class	Sample (n)	Total trees in the plots	Bark thickness (mm)	Bark WD (g cm ⁻³)	Core WD (g cm ⁻³)	Weighted average WD (g cm ⁻³)
< 0.350	10	23	10.1 ± 5.2	0.371 ± 0.084	0.295 ± 0.041	0.296 ± 0.039
0.351-0.500	31	107	7.9 ± 4.0	0.529 ± 0.177	0.431 ± 0.035	0.432 ± 0.034
0.501-0.650	171	619	5.8 ± 3.5	0.552 ± 0.194	0.595 ± 0.038	0.595 ± 0.038
0.651-0.800	315	1162	5.5 ± 3.1	0.620 ± 0.185	0.717 ± 0.040	0.717 ± 0.040
0.801-0.950	134	699	4.5 ± 2.4	0.719 ± 0.208	0.865 ± 0.046	0.865 ± 0.046
> 0.951	19	54	5.3 ± 2.0	0.675 ± 0.184	0.985 ± 0.033	0.983 ± 0.034
Total (mean ± SD)	680	2664	5.6 ± 3.3	0.616 ± 0.201	0.704 ± 0.134	0.703 ± 0.133

each plot, here defined as the arithmetic mean of wood density (weighted between bark and core wood) using as reference the values (measured and estimated) of all tree individuals present in each plot. To obtain the specific average of each plot, we took the values of the species sampled for wood density (25% of the individuals) and replicated them for the other individuals of the same species present in the plot that were not sampled for wood density. The wood density of the species not sampled in the plots was resolved using the values obtained for the same species in the dominant forest type (ombrophilous, semideciduous or deciduous). When not available, we adopted the genus or family average – first within the plot and then by the dominant forest type.

The specific average of each plot represents the functional composition of the tree community at the species level. Therefore, differences in this mean between communities mix the effect of differences in species composition with that of intraspecific variation (e.g., due to phenotypic plasticity or local adaptation). To separate these sources of variation, we calculated the fixed average per plot, defined using mean wood densities of the species occurring in one plot, weighted by their respec-

tive relative abundances. The fixed average can only vary between plots if the species composition changes, and it thus represents interspecific variation, explaining the effect of species-substitution (turnover). Consequently, the difference between the specific average and the fixed average represents intraspecific variation (Lepš et al. 2011), which explains the variation in the wood densities of the sampled individuals of a given species.

The proportion of total variation represented independently by the two community composition components (inter- and intra-specific variation), as well as covariation between them, was estimated using an ANOVA of the median densities recorded for wood (Lepš et al. 2011). This only measures the relative contribution of inter- vs. intraspecific variation to changes in average wood density across plots but does not explain why the values of this average change, nor does it indicate how each of these two components responds to environmental variation. Accordingly, each of the two components was used separately as a dependent variable in a multiple-regression model, with environmental variables as predictors: (i) altitude (m a.s.l.); (ii) available phosphorus content; (iii) sum of bases (K + Ca + Mg); (iv) sum of

micronutrients (Fe, Zn, Mn, Cu and B); (v) soil clay percentage (%); (vi) soil organic matter content; and (vii) pH. To visualize statistically significant effects, we used conditional graphs that showed variation in the dependent variable in relation to the primary predictor variable, using partial residuals (Breheny & Burchett 2017). All analyses were performed using the R computational platform, ver. 3.6.3 (R Core Team 2020) taking into account values specified in Tab. S1 (Supplementary material).

Results

Data description

The weighted average wood density (± standard deviation, SD) for the tree assemblage in the ecotone forest in the eastern portion of Maracá was 0.703 ± 0.133 g cm⁻³ (range: 0.203-1.102 g cm⁻³ – Tab. 1, Tab. S2 in Supplementary material). The distribution of the wood density values of most of the individuals (69%) that we randomly sampled in the fieldwork was observed for the categories with WD > 0.650 g cm⁻³ (medium-hard = 315; hard = 134; very-hard = 19 – Fig. 2a). This sample distribution has a direct relation to the distribution of the total population of individuals present in all sampled plots (Fig. 2b).

The species with the lowest wood densities were *Schefflera morototoni* (Araliaceae, 0.324 ± 0.012 g cm⁻³) and *Apeiba tiburou* (Malvaceae, 0.346 ± 0.111 g cm⁻³ – Tab. S2), both generally associated with well-drained plots (> 65 m a.s.l.). *Peltogyne paniculata* (Leguminosae, 0.921 ± 0.032 g cm⁻³) and *Peltogyne gracilipes* (Leguminosae, 0.902 ± 0.089 g cm⁻³) were the species with the highest wood densities, and they almost always occurred in plots located in seasonally flooded environments (≤ 65 m a.s.l.).

Wood density vs. environmental variables

Altitude (a proxy for drainage), clay content and sum of micronutrients in the soil explained 23% of variation in mean wood

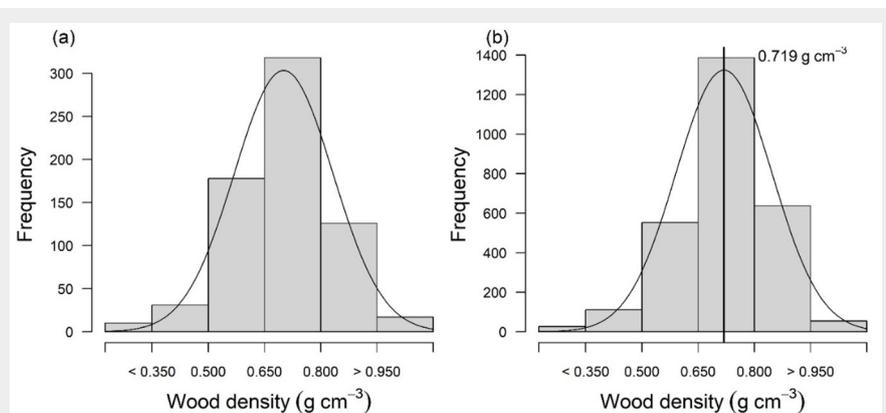


Fig. 2 - Frequency distribution of wood density for (a) samples (n = 680) and (b) total individuals (n = 2768).

density at the community level (Tab. 2). Regression models indicate negative relationships with altitude (Fig. 3a) and soil clay content (Fig. 3b), and a positive relationship with the sum of micronutrients (Fig. 3c). Mean wood density in seasonally flooded plots (≤ 65 m a.s.l.; $n = 32$; 0.742 g cm^{-3}) was higher ($t_{0.05} = 2.018$; $p = 0.0137$) than that for plots in flooding-free areas (> 65 m a.s.l.; $n = 97$; 0.713 g cm^{-3}). This result indicates that average wood density of tree communities occurring in seasonally flooded areas (low altitude) with low clay content (sandy soils) and high sum of micronutrients (especially those related to flooding periodicity) tend to be higher than for tree communities in flooding-free areas. There was no detectable spatial autocorrelation in the residuals of our models and, thus, no need for more-complex models (Fig. S1 in Supplementary material).

Intra- and inter-specific variability in wood density

Separation of the variation in wood density into components explained by species-substitution (turnover) and intraspecific variation revealed that substitution explained most of the total variation (86%), with ~13% of the variation attributed to intraspecific variation for the wood densities sampled (Fig. 4). The species-substitution component responded similarly to the patterns previously observed for environmental variables (hydro-edaphic and topographic conditions). Thus, altitude, clay content and sum of micronutrients were slightly better predictors in the case of species substitution and explained 26% of the variation in wood density (Tab. 2). By contrast, phosphorus content explained 15% of the intraspecific variation in wood density (Fig. 5), so that individuals occurring in more phosphorus-rich soils tend to have lower wood density.

Discussion

Biases in wood density estimates

WD estimates are subject to various kinds of biases. In our study we used samples

Tab. 2 - Regression models relating the variation in wood density to drainage (altitude), and physical and soil chemical properties in ecotone forests in northern Amazonia ($n = 129$). (SOM): soil organic matter; (**): $p < 0.01$; (*): $p < 0.05$.

Response	R ²	Predictor	Coefficient	t	P
Mean wood density	0.23	Intercept	0.8535	-	-
		Altitude	-0.0015	-2.533	0.013**
		Clay content	-0.0028	-2.382	0.019**
		pH	-0.0036	-0.189	0.850
		SOM	0.0012	1.282	0.202
		Available P content	-0.0061	-1.612	0.109
		Sum of bases	-0.0251	-1.502	0.136
		Soil micronutrients	0.0001	2.110	0.037*
Mean wood density due to species turnover	0.26	Intercept	0.7856	-	-
		Altitude	-0.0015	-2.816	0.006**
		Clay content	-0.0023	-2.179	0.031*
		pH	0.0139	0.805	0.422
		SOM	0.0008	0.989	0.325
		Available P content	-0.0029	-0.827	0.410
		Sum of bases	-0.0204	-1.323	0.188
		Soil micronutrients	0.0002	2.685	0.008**
Mean wood density due to intraspecific variation	0.14	Intercept	0.0679	-	-
		Altitude	-0.0000	-0.142	0.887
		Clay content	-0.0004	-1.018	0.310
		pH	-0.0104	-1.470	0.144
		SOM	0.0003	0.999	0.320
		Available P content	-0.0032	-2.278	0.024*
		Sum of bases	-0.0048	-0.767	0.445
		Soil micronutrients	0.0000	0.952	0.343

taken at breast height using an increment borer, which allows living trees to be sampled without felling them. Each sample consists of a cylindrical core taken from the bark to the center of the trunk. In our study we separated the bark from the core wood and used a weighted average of their respective densities based on a linear fraction (radius) of the stem cross-sectional represented by each, thus eliminating one of the biases inherent in this method. However, the core wood was not sepa-

rated into its components (sapwood and heartwood). The heartwood is over-represented in the sample because, being located in the center of the trunk, each centimeter of the core in the heartwood corresponds to a smaller-diameter circle than a centimeter in the sapwood portion of the core. In almost all Amazonian trees the heartwood is denser than the sapwood (in contrast to many temperate-zone trees), resulting in an upward bias in the estimated wood density. A rough estimate of

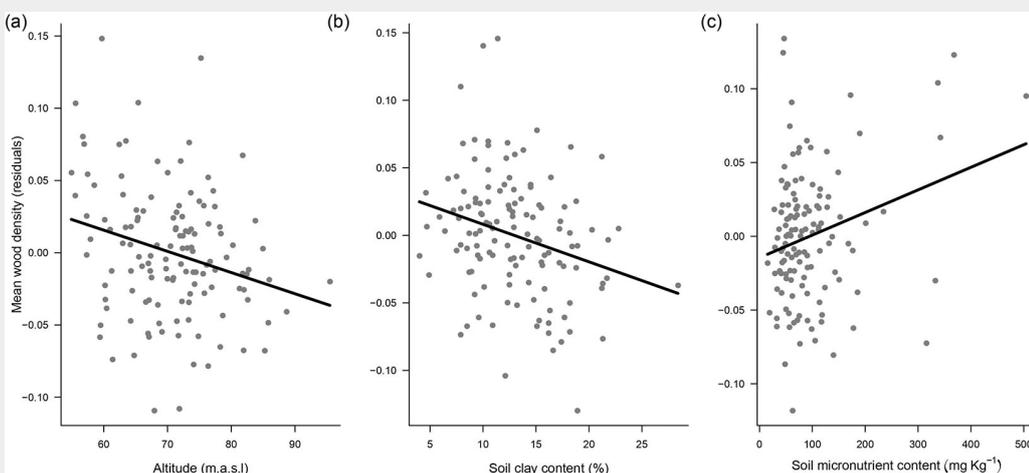
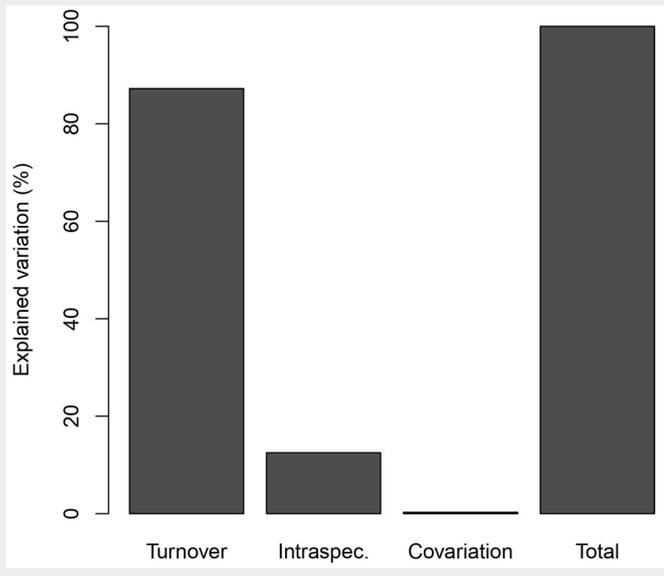


Fig. 3 - Mean wood density ($n = 129$ plots) vs. (a) altitude, (b) clay content and (c) soil micronutrients.

Fig. 4 - Decomposition of the variation of wood density into components of species substitution and intraspecific variation.



Wood density vs. environmental variables

Our study relating environmental conditions to intra- and inter-specific spatial variation in WD is the first to deal with tree species assemblages that occur in ecotone forests in the northern Brazilian Amazon. The results indicate that average WD of tree communities occurring in seasonally flooded areas (shallow water table) with low clay content (sandy soils) and high sum of micronutrients (especially those related to flooding periodicity) tend to be higher than those tree communities in flooding-free habitats. Our investigation revealed that environmental characteristics explain differences between intra- and inter-species values, as previously observed in other regions of the Amazon (Baker et al. 2004, Muller-Landau 2004, Chave et al. 2006, Poorter et al. 2019). However, we also separated community-based spatial variation in WD into categories of species substitution and intraspecific variation, revealing their divergent responses to environmental variation, and suggesting the presence of overlooked patterns in tree adaptive response to hydro-edaphic variation. This integrated analytical approach helps improve understanding of how variation in WD occurs spatially, reducing the degree of uncertainty in an especially important variable for calculating carbon and biomass stocks in Amazonian forest ecosystems.

In general, the values we determined for each species (range: 0.203-1.102 g cm⁻³) lay within the ranges of values cited in databases for tropical-forest species (Chave et al. 2006, Zanne et al. 2009). However, our community-averaged mean wood density (0.719 cm³) is 12% higher when compared with the general average value presented by Nogueira et al. (2007) for forests in the Brazil's "arc of deforestation" (0.642 cm³), or 7.3% (0.67 g cm³) to 23.9% (0.58 g cm³) higher when compared to the results obtained by Baker et al. (2004) in different Amazonian regions. These differences indicate a relation to the intrinsic characteristics of the studied ecotone forests, especially the higher hydro-edaphic restrictions that characterize some lowlands areas on the eastern portion of Maracá. Most of these areas are seasonally flooded (shallow water table) where poor sandy soils (low clay content) associated with large sums of Fe, Zn, Mn and Cu predominate. Limitation of forest productivity from toxicity caused by either micronutrient deficiency or by excessively high concentrations of micronutrients is not common in tropical soils (Binkley & Fisher 2019). However, higher availability of micronutrients associated with waterlogged tropical soils can reach toxic levels for some plants, conferring more restricted aspects to the habitat (Davies 1997). In addition, areas in the eastern portion of Maracá with higher environmental restrictions are generally populated by *Peltogyne gracilipes* (Nascimento et al. 2017, Villacorta et al. 2022), a mon-

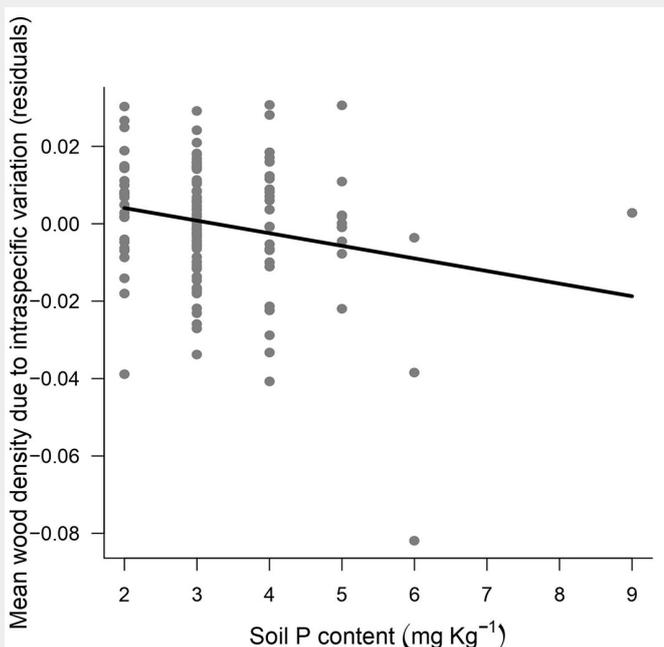
the magnitude is provided by information from other parts of Brazilian Amazonia (Fearnside 1997). Based on 57 species from sites near Santarém in the state of Pará, an average of 9.7% of the commercial volume is sapwood, and, based on 33 species near Manaus, Amazonas, the density of the sapwood averages 94.7% of the density of the corresponding heartwood; these values together imply a 0.52% upward bias. The amount of bias varies with the diameter distribution of the forest and with forest type, for example between central and southern Amazonia (Nogueira et al. 2008).

Another bias in the same direction results from the samples being taken only at the DBH height, as wood density decreases along the length of the trunk. For example, based on 12 species near Manaus, the average wood density at the midpoint of the commercial bole (the trunk to the first sig-

nificant branch) is 7.1% lower than the density at the DBH height (Fearnside 1997). The relationship between density at the DBH height and that of the entire bole varies with forest type, as between central and southern Amazonia (Nogueira et al. 2008).

These biases affect virtually all estimates of WD made based on cores from increment borers, such as those from important network of plots surveyed by the Biodiversity Research Program (PPBio: <https://ppbio.inpa.gov.br/>) that includes the Maracá plots used in the present study. These biases affect estimates of carbon stocks and of greenhouse-gas emissions from deforestation that are based on WD data from increment-borer cores but do not affect our conclusions on the importance of intra- and inter-specific variability in driving the divergences in woody biomass.

Fig. 5 - Intra-specific variation in wood density versus soil phosphorus content in ecotone forests in the eastern portion of Maracá Island.



odominant species characterized by high wood density (0.901 g cm^{-3}).

The relationship between WD and more/less restricted environmental characteristics has already been commented on for other regions of the Amazon (Parolin & Worbes 2000, Baker et al. 2004, Wittmann et al. 2006). This relationship also holds in the case of Maracá, since the species with highest abundances in the locality (*P. gracilipes*, *Ecclinusa guianensis* and *Lecythis corrugata* – see Silva et al. 2019) have high WD values and occur preferably in ecotone-forest areas with strong environmental restrictions. Our results are different from those of Costa et al. (2023), who reported that lower WD values are characteristic of environments with shallow water tables (seasonally flooded). However, our findings were similar to those reported by Souza (2014) in the Viruá National Park (Roraima), a forest formed by contact between seasonally flooded oligotrophic ecosystems and open forests with poor soils, where a set of sampled trees had a higher mean WD (0.700 g cm^{-3}) when compared with other Amazonian regions. This environmental conditioning, where drier regions in northern Brazilian Amazonia can transform seasonally flooded areas into zones with enough moisture to support a strong dry season and induce plant production with slow growth, can promote species with higher wood density values (e.g., *P. gracilipes*). Our findings imply that, in ecotone forest areas with high environmental restrictions in northern Amazonia, a group of a few species with both higher WD and higher abundance can lead to a higher community-averaged WD, considering the entire tree assemblage.

Intra- and inter-specific variability in wood density

Our results showed that, in tree assemblages occurring in ecotone forests in the eastern portion of Maracá, environmental variables such as altitude, clay content and soil micronutrients determined most of the variation in wood density, with about 13% of the explained variation being attributed to intraspecific variation in wood density. The negative relationship between the altitudinal gradient and wood density differs from results in the central Amazon region reported by Cosme et al. (2017). These authors found that altitude controls the availability of water in the soil, and, in response to water stress, plants develop physiological adaptations (e.g., smaller average stem diameter, smaller average stem area and less sapwood area), resulting in high-density wood for improved hydraulic tolerance at higher altitudes (Jucker et al. 2018, Oliveira et al. 2019). However, in our study area the physical and soil chemical properties are acting in the opposite way.

We found species with the highest wood density values in seasonally flooded areas with strong hydro-edaphic restrictions (e.g., poorly drained soils with temporary

Tab. 3 - Differences in the aboveground live biomass (Mg ha^{-1}) in ecotone forests of the eastern portion of Maracá Island considering new wood density values for the study area. (FWP): forest without *Peltogyne* (flooding-free areas located on flat relief); (PPF): *Peltogyne*-poor forest (slopes occasionally covered by rocky fragments); (PRF): *Peltogyne*-rich forest (poorly drained soils - seasonally flooded), following previous values and definitions of Nascimento et al. (2014).

Types	Previous values (Nascimento et al. 2014)	Current values (this study)	Change in stand biomass (%)
FWP	363.35	368.29	+ 1.36
PPF	433.88	462.65	+ 6.63
PRF	422.97	492.14	+ 16.35

anoxia), indicating that habitat heterogeneity is selecting species with phenotypic traits adapted to local environmental conditions (Muller-Landau 2004, Swenson & Enquist 2007). This result indicates that most tree species adapted to seasonally flooded areas (temporary anoxia) in the ecotone forests of Maracá tend to have lower growth rates due to an environment characterized by stressful edaphic conditions, which explains their higher wood-density values. The relation between temporal anoxia, lower growth rates and higher wood density values cannot be applied in a general way to other types of forests in the Amazon that have environmental characteristics similar to those analyzed in the eastern portion of Maracá. This could produce errors in biomass and carbon calculations that would be difficult to correct.

Most of the between-plot variation was due to changes in species composition, where plots located on seasonally flooded areas have mean wood density 4% (0.742 g cm^{-3}) higher than those in flooding-free areas (0.713 g cm^{-3}). In addition, when we isolated only intraspecific variation, wood density appears to depend on soil phosphorus levels. Condit et al. (2013) indicated that phosphorus availability in soils, which is a limiting resource in Amazonian forests, is primarily responsible for the distribution of species on environmental gradients. Species that have a wide distribution along the phosphorus gradient may have genetically specialized populations in P-poor habitats as part of an evolutionary strategy to improve the plant's hydraulic architecture and increase stem longevity (Oliveira et al. 2019). On the other hand, Zalamea et al. (2016) suggested that phosphorus-dependent growth rates provide an additional explanation for the distribution of pioneer tree species with high phosphorus demand; these are species that naturally have lower wood density. The results at Maracá agree with this logical tendency, where the most-restricted habitats have trees that grow more slowly and produce denser wood. At Maracá the habitats with hydro-edaphic restrictions (seasonally flooded areas) also have poor soils characterized by low levels of phosphorus, and the main

species in this habitat (*P. gracilipes*) has high wood density and a slow growth rate.

Finally, our study advances the understanding of factors determining wood-density variability in ecotone forests in the northern Brazilian Amazon, allowing to improve estimates of woody biomass and carbon stocks in this poorly studied region. For example, we recalculated the aboveground live biomass estimated in Maracá by Nascimento et al. (2014) using our wood-density dataset (Farias et al. 2020) instead of the data used by the authors from global repositories (Zanne et al. 2009). We calculated that the above-ground live biomass values presented by these authors were underestimated by from 1.4% to 16.3% for the main forest types occurring in Maracá (Tab. 3). This local example shows that, despite the recent advances related to data on wood density, it is still necessary intensify our understanding of how environmental filters determine spatial patterns of wood density in the forests of the northern Brazilian Amazon. This advance is needed to improve the current estimates of carbon and biomass stocks in the Amazon because, in general, wood-density data are derived from databases generated in the eastern, western, or central Amazon or in other tropical regions of the world. Biomass is proportional to wood density, and, because of the vast areas of Amazonian forests, small percentage differences in these estimates correspond to very large amounts of carbon.

Conclusion

Wood density variation between tree communities in ecotone forests studied in Maracá Island is partially driven by hydro-edaphic conditions. Overall, our results indicate that most of the between-plot variation in wood density was due to changes in species composition. Analysis of interspecific variability indicates that forests occurring in environments with higher hydro-edaphic restrictions (e.g., seasonally flooded soils) contain species with higher wood density than those environments with lesser restrictions (e.g., non-flooded soils). Our study advances the understanding of factors determining the variability of tree wood density in regional ecotone forests,

allowing improvement of estimates of the woody biomass and carbon stocks in the forests of the northern Brazilian Amazon.

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Supplementary Material

Fig. S1 - Spatial correlograms for our three multiple regression models, created with function `spline.correlog()` from the R package "ncf".

Tab S1 - Database used in the analysis of intra- and inter-specific variations in wood density in ecotone forests of the eastern portion of Maracá Island, Roraima State, northern Brazilian Amazon.

Tab. S2 - Tree species and morphospecies wood density estimate for ecotone forests in the eastern portion of Maracá Island, northern Brazilian Amazonia (mean \pm SD).

Link: [Farias_4137@suppl001.pdf](#)