

Landsat TM imagery and NDVI differencing to detect vegetation change: assessing natural forest expansion in Basilicata, southern Italy

Giuseppe Mancino, Angelo Nolè, Francesco Ripullone, Agostino Ferrara

The NDVI (Normalized Difference Vegetation Index) differencing method using Landsat Thematic Mapping images was implemented to assess natural expansion of forests in the Basilicata region (southern Italy) for the period 1984 through 2010. Two Landsat TM (Thematic Mapper) images (1984-2010) were georeferenced and geographically corrected using the first order polynomial transformation, and the nearest neighbour method for resampling. The images were radiometrically corrected using the dark object subtraction model. The pre-processed Landsat TM images were used to calculate NDVI, and subsequently for NDVI differencing. Finally, a threshold for vegetation change detection was identified by visual analysis of Landsat TM RGB band composition, and ratios and visual comparison of digital aerial orthophotos. The methodology was validated using ground-truth observations over the study area. The applied method showed 91.8% accuracy in detection of natural forest expansion. During the examined period, total regional forest cover increased by 19.7% (70 154 ha), consistent with National Forest Inventory data (1984-2005). The observed forest expansion was also examined in relationship with landscape physical characteristics and distribution of vegetation types in the Basilicata region. Surprisingly, considerable forest expansion also occurred on degraded soils in drought-prone Mediterranean areas.

Keywords: NDVI Differencing, Landsat TM, Detection Change, Natural Forest Expansion

Introduction

Assessment of natural forest expansion represents a crucial issue to elucidate several processes, including biogeochemical cycles, atmospheric composition related to climate change, and forest carbon uptake, as well as socio-economic processes and issues. Anthropogenic and naturally induced land cover changes affect spatial and temporal distribution and availability of environmental resources, and alter ecosystem composition and productivity. Globally, these processes can be considered the primary catalysts for change in biogeochemical cycling, atmospheric composition, and climate (Pielke

2005, Metz et al. 2007, Turner et al. 2007). Forest land-use and land-cover change (LU-LCC) were recognised as key issues in greenhouse gas removal/emission processes as specified by the Good Practices Guidance for Land Use, Land Use Change, and Forestry (GPG-LULUCF) during the Intergovernmental Panel on Climate Change (IPCC) established at the Kyoto Protocol (Penman et al. 2003). Observation and assessment of forest cover changes are crucial to elucidate the complexities inherent in feedback processes between forest distribution and human activities in sustainable forest development, natural resource management, biodi-

versity conservation, ecosystem functioning, and biogeochemical cycling (IPCC 2007). In Mediterranean regions, natural forest expansion is primarily related to the abandonment of agricultural practices and cattle-raising in marginal areas representing the principal change in Italy's Mediterranean rural landscape over the past five decades (Piussi 2005). These processes generally vary in terms of the vegetation successional series and time scale, however the expansion dynamics are shared, beginning with an initial phase of spontaneous shrub dominance, followed by tree colonisation (Biondi et al. 2006).

In recent decades, satellite-based high-resolution observations with multispectral scanners provided the scientific community with consistent data to implement detailed thematic mapping for local and regional scale land classification (Friedl et al. 2002, 2010, Lu & Weng 2007, Giri 2012). The near infrared wavebands on the Landsat Thematic Mapper (TM) facilitates advanced land classification analyses based on differences in spectral reflectance of different land cover types. In particular, specific foliar reflectance, pigment absorptions, and foliar moisture wavelength ranges represent the basis of vegetation class analyses. Furthermore, the availability of such land cover data at different spatial and temporal scales promotes the development and implementation of vegetation change detection techniques, which furthers our understanding on vegetation and ecosystem dynamics (Cohen & Fiorella 1998, Coppin et al. 2004, Lu et al. 2004, Martinez & Gilabert 2009).

In forest ecosystems, land cover change dynamic detection based on visual and statistical approaches represents a challenge to the scientific community due to the difficulties in remotely sensed image acquisition errors as remnant geometric errors, atmospheric scatter, and cloud effects. Other main sources of error are related to physical variability in land characteristics, including topography, vegetation types, and phenology (Shoshany 2000, Roy et al. 2002). In particular, Verbesselt et al. (2010) proposed categorising vegetation changes at the ecosystem level under different classes based on a temporal scale (seasonal, gradual, and abrupt changes), and on the factors driving change, such as seasonal environmental constraints, inter-annual climate variability, land management, and disturbances (deforestation, fires, and floods). In many cases, land cover changes in highly fragmented forest areas can be mismatched with inter-annual variability due to plant phenology or seasonal productivity patterns related to climate variability (De Beurs & Henebry 2005). The comparison between the time-series of remotely sensed imagery might include a combination

□ School of Agricultural, Forestry, Food and Environmental Sciences, University of Basilicata, v. N. Sauro 10, Potenza (Italy)

@ Angelo Nolè (nole@unibas.it)

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Fig. 1 - Study area: Basilicata Region, southern Italy.



from other surfaces based on green vegetation chlorophyll absorption of red light for photosynthesis, and reflection of NIR wavelengths (Tucker 1979). The NDVI differencing technique was widely applied for both human-induced and natural forest cover change detection as land cover conversion, forest harvesting, revegetation, or afforestation that includes natural forest expansion and human-induced landscape restoration (Lyon et al. 1998, Wilson & Sader 2002, Sader et al. 2003, Lunetta et al. 2002, 2006, Nabuurs et al. 2007, Poteh et al. 2009). In addition to the standardized techniques for pre-processing and differencing images, the most important step for vegetation change detection analysis is discrimination between real changes and seasonal or inter-annual variability, represented by a threshold between these factors, which is generally determined by applying the standard deviation (SD) from the NDVI differencing image (Hayes & Sader 2001, Coppin et al. 2004, Desclée et al. 2006, Sepehry & Liu 2006, Pu et al. 2008).

In the present study, we propose a vegetation change detection analysis based on the NDVI differencing technique to assess forest cover changes related to natural forest expansion for the Basilicata region (southern Italy) from 1984 through 2010. The objectives were as follows: (i) to develop a procedure for pre-processing Landsat TM imagery, NDVI differencing, and identification of a threshold for vegetation change detection; and (ii) to determine primary patterns of the natural forest expansion in Mediterranean environments.

Material and methods

Study area

The study area represents the Basilicata region in southern Italy (Fig. 1). The regional territory is dominated by a mountainous area to the west, flanked by the Apennine Mountain range, while the east is comprised of low hills and wide valleys grading into flat plains on the south coast, along the Ionian Sea. The topography is consistent with climate characteristics, annual temperatures, and precipitation patterns indicating a Mediterranean humid type in the Apennine Mountain range, and a Mediterranean dry climate type in the hilly and flat areas. The annual rainfall average recorded at Senise and Terranova di Pollino was respectively 730 mm, and from 416 mm to 1211 mm, and annual temperatures ranged from a mean minimum temperature of 10 °C to mean maximum temperatures of 20-24 °C (Cantore et al. 1987, Fiorenzo et al. 2008).

Forests cover 35.6% (355 367 ha) of the total regional surface (999 400 ha - ISTAT 2001). The Regional Forest Atlas (Costantini et al. 2006) reported deciduous oak forests

of seasonal variability and abrupt changes; therefore a long-term dataset represents the fundamental choice for change detection (Setiawan & Yoshino 2012). Several proposed vegetation change detection methods are based on the same image pre-processing to create a time-series dataset, requiring a geometric and radiometric image correction. Coppin & Bauer (1996) and Milne (1988) reported the main methodological approaches for vegetation change detection can be distributed into four broad categories: (i) linear procedures (difference and ratio images); (ii) classification routines (post-classification change, spectral pattern change); (iii) transformed data sets (vegetation indexes, principal components analysis-PCAs); and (iv) others, such as regression analysis, knowledge-based expert systems, or neural

networks (Sader et al. 2003).

Several literature reviews (Muchoney & Haack 1994, Nordberg & Evertson 2004, 2005) reported the most efficient methodologies in accuracy and cost saving performances were image differencing and PCA techniques. The image differencing technique is based on a cell-by-cell subtraction between different images in a time-series. This technique applies differences between remotely sensed images of vegetation characteristics, and indexes derived from image radiance or reflectance differencing, *i.e.*, NDVI (Normalized Difference Vegetation Index) or change vector differencing (Schowengerdt 1997). The NDVI differencing method uses estimated NDVI as the normalized difference between near infrared (NIR) and visible red (RED) bands, which discriminate vegetation

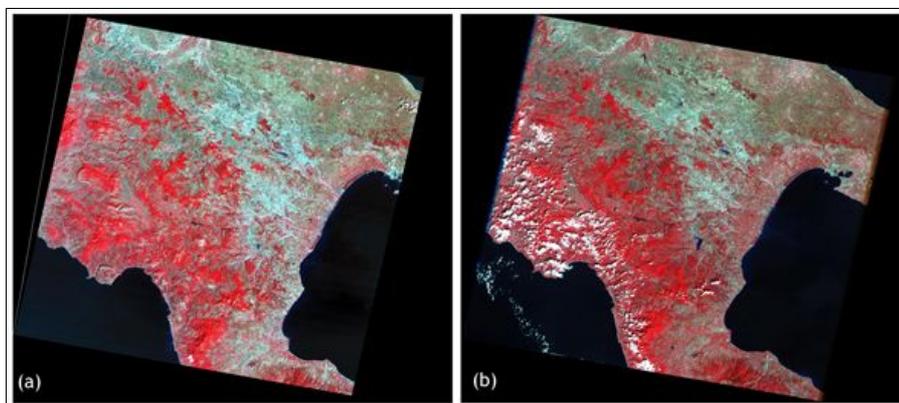


Fig. 2 - Landsat TM images (Path 188, Row 32) Band composition 4-3-2; image dates: (a) 08 September 1984; and (b) 16 September 2010.

representing 51.8% of the forest cover, while other forest types were generally characterized by lower extensions, including Beech forest (8.4%), Mediterranean Macchia (7.9%), shrublands (6.9%), and other deciduous broad-leaved forests (5.5%). The remaining 19.5% were comprised of evergreen broad-leaved forests, Mediterranean pine forests, Chestnut stands, hygrophilous forest, and Garrigue. Coppices represented 51.6% of the total managed forests, while high stands represented the remnant.

Image pre-processing

Two Landsat TM cloud free images (path 188, row 32) with 30 x 30 m spatial resolution were analysed (acquiring image periods: 8 September 1984 and 16 September 2010 - Fig. 2). Image registration and/or rectification are required to facilitate image conformity to another image, and involves georeferencing if the reference image is already rectified to a particular map projection. Landsat TM 2010 was projected using 26 ground control points (GCPs) to the Universal Transverse Mercator (UTM) projection System (zone number: 33N; reference datum: WG-S84). The first order polynomial transformation model and nearest neighbor method for resampling were used for geographical correction of the two images, with a Root Mean Square (RMS) error of 0.5 pixels. Subsequently, the TM 1984 image was geo-coded to the TM 2010 image by the "map-to-map" method, and resampled with a first-order polynomial nearest neighbor algorithm using 12 GCPs for the registration process. The Root Mean Square Error (RMSE) for the map registration process was 0.5.

The two images were subsequently clipped to the final study area, and radiometrically corrected using the dark object subtraction model (Chavez 1996), a widely applied methodology considered one of the best approaches for the radiometric correction in change detection analysis (Song et al. 2001, Lu et al. 2002, Hu et al. 2004, Schroeder et al. 2006, Mancino et al. 2009, Wang & Xu 2010). The reliability of the radiometric correction was confirmed using an inter-calibration algorithm between the NDVI images over a test area of the study site. Among the numerous radiometric normalization approaches specific to Landsat data (Elvidge et al. 1995, Jensen 1996, Yuan & Elvidge 1996, Yang & Lo 2000, Callahan 2001, Over et al. 2003, Mateos et al. 2010), we selected the inter-calibration algorithm, defined as no-change (NC) regression normalization, which has already been applied over areas characterized by very complex landscape patterns (Simoniello et al. 2008). Comparisons of results between the dark object subtraction model and the inter-calibration algorithm of NC regression normalization showed no significant differences in change detection over

forest areas between the two methods. Cross-classification between the two images exhibited a 2% difference for analysis of all land use changes, which was reduced to less than 1% for the assessment of only forest land use changes.

NDVI differencing

Change detection in land use alterations during the analysis period (1984-2010) was analyzed using the NDVI differencing technique following Singh (1989). First, we calculated NDVI following the general normalized difference between Band TM4 (near infrared - NIR) and band TM3 (visible red - RED) from the two Landsat TM pre-processed images (1984 and 2010 - eqn. 1):

$$NDVI = \frac{TM4 - TM3}{TM4 + TM3}$$

The resulting images were subtracted to assess the $\Delta NDVI$ image with positive (NDVI increase) and negative (NDVI decrease) changes on a 30 x 30 m pixel resolution (eqn. 2):

$$\Delta NDVI = NDVI_{2010} - NDVI_{1984}$$

The NDVI difference image was also tested to determine its goodness-of-fit to a normal distribution. Mean, mode, median, standard deviation, and specific statistical indexes

were generated, including skewness (Kendall & Stuart 1969, Groeneveld & Meeden 1984), and kurtosis (Balanda & MacGillivray 1988), and Kolmogorov-Smirnov non-parametric tests (Lilliefors 1967, Justel et al. 1997) were conducted.

The difference image $\Delta NDVI$ was then reclassified using a threshold value calculated as $\mu \pm n \cdot \sigma$; where μ represents the $\Delta NDVI$ pixels digital number mean, and σ the standard deviation. The threshold identifies three ranges in the normal distribution: (a) the left tail ($\Delta NDVI < \mu - n \cdot \sigma$); (b) the right tail ($\Delta NDVI > \mu + n \cdot \sigma$); and (c) the central region of the normal distribution ($\mu - n \cdot \sigma < \Delta NDVI < \mu + n \cdot \sigma$). Pixels within the two tails of the distribution are characterized by significant vegetation changes, while pixels in the central region represent no change. The n factor defines the range of dispersion around the mean. This study considered only the positive variation in forest cover defined as the area of probable natural forest expansion.

Threshold identification for detection of vegetation changes represents a key issue in the NDVI differencing method. The standard deviation (σ) is one of the most widely applied threshold identification approaches for different natural environments based on different remotely sensed imagery (Singh 1989, Jensen 1996, Coppin et al. 2004, Hu et al. 2004, Lu et al. 2004).

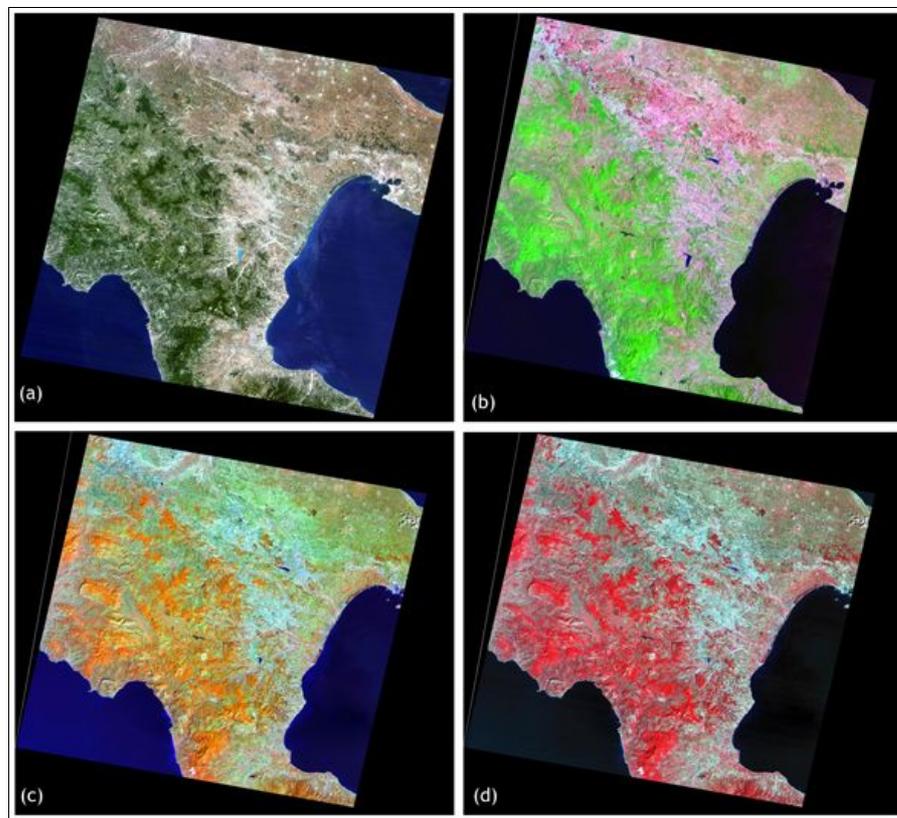
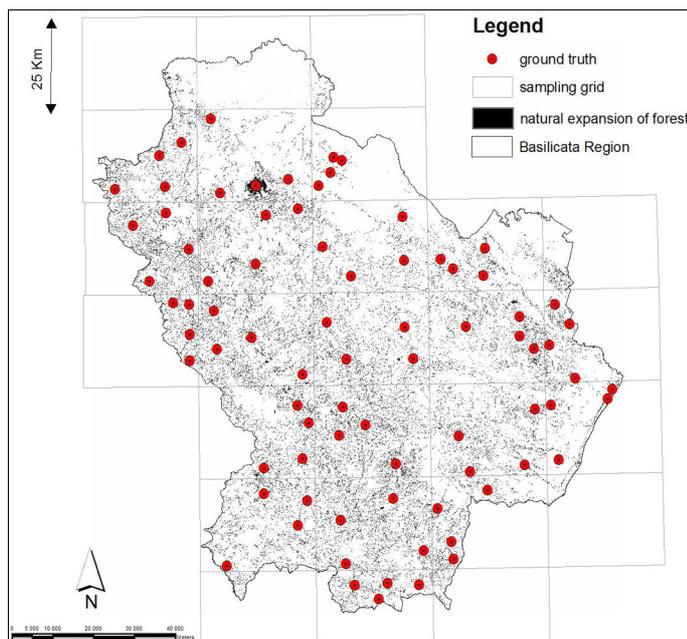


Fig. 3 - Landsat TM band composition. (a) TM3-2-1; (b) TM7-4-2; (c) TM4-5-1; (d) TM4-3-2.

Fig. 4 - Ground-truth points sampling scheme.



Generally, the threshold value is identified by $n\sigma$ of the NDVI difference image average, where the n value is identified by the trial and test method, and σ is the standard

Tab. 1 - Parameters of the distribution of the NDVI difference image values. (**): $p < 0.01$.

Parameter	Value
Mean	0.01
Median	0.01
Mode	0.011
Standard Deviation	0.101
Skewness	-0.127
St. Error of Skewness	0.025
Kurtosis	1.927
St. Error of Kurtosis	0.049
KS Test	0.0756**

deviation of the pixel values density function in the change image. This approach exhibited viable results, and reliability for different forest ecosystems under both human-induced and natural land use changes, with threshold values between 1σ and 2σ supported in the literature (Yuan et al. 1998, Mas 1999).

In the present study, the final identification of the best-fitting $n\sigma$ threshold value was based on visual analysis of Landsat TM RGB band composition and ratios, and on the visual comparison of digital aerial orthophotos *Volo Italia 1994* (black and white) and *Terraitaly 2008* data sets. In particular, a visual analysis based on three different threshold values (1σ , 1.5σ and 2σ) was conducted over 200 points randomly chosen within areas of vegetation change using the 1σ threshold, in order to represent the areas with the most probable vegetation changes.

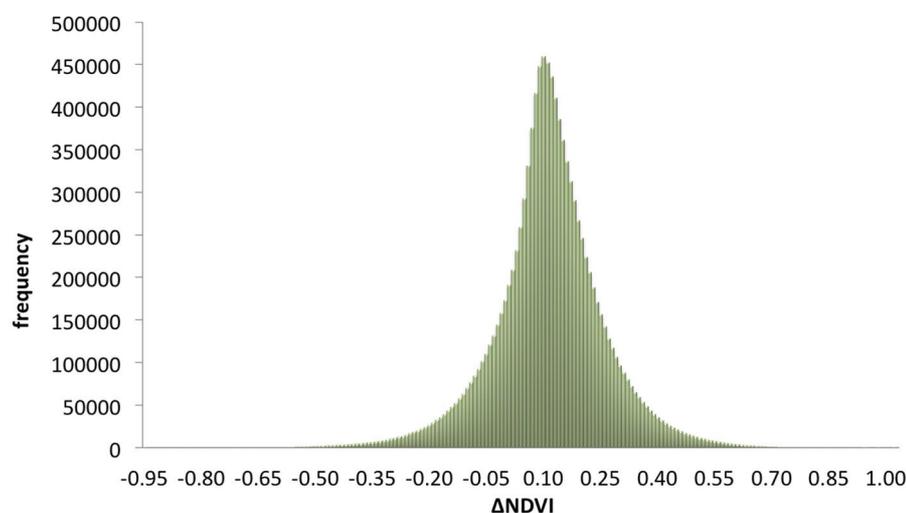


Fig. 5 - Distribution of Δ NDVI pixel values.

Landsat TM RGB band compositions were selected (Fig. 3) since their efficient representation of the vegetation cover (Lillesand & Kiefer 1999, Gibson & Power 2000, Horning et al. 2010). Landsat TM4-3-2 is commonly used for the identification of broad leaf, conifers, grasslands, sparsely vegetated areas and/or healthier vegetation (Jensen 2000, Dehnavi et al. 2010). Landsat TM4-5-3 is used for the discrimination of vegetation types and conditions and land/water interface (Jensen 1996). The composition Landsat TM4-5-1 is particularly efficient in the detection of healthy vegetated areas, but also for recently clear-cut regions and new vegetative growth generally attributed to sparse grasslands (Short 1982, Lillesand & Kiefer 1999). Landsat TM7-4-2 was analyzed for the detection of low-density vegetation that can represent areas with probable natural expansion due to forest processes (Richards & Xiuping 2006). The composition Landsat TM5-4-1 representing healthy vegetation and is generally used for the detection of agricultural areas (Wallace et al. 2006). In addition to the RGB band combinations, the Landsat TM band ratios were analyzed using the Landsat TM4/TM3 and TM3/TM2. The TM4/TM3 ratio distinguishes vegetation, water, and croplands, while the TM3/TM2 ratio separates forests and croplands, which is useful in discriminating broad classes of vegetation (Sabins 1997). Finally, the $n\sigma$ threshold value was statistically determined by Cohen's Kappa inter-rater agreement coefficient (Cohen 1960, Story & Congalton 1986, Congalton 1991, Jenness & Wynne 2005), which identified $n = 1.5$.

Ground-truth validation

Natural expansion of forest areas within positive vegetation change was detected by clipping the $\text{NDVI} > +1.5\sigma$ image using a mask image of the Forest Map for Basilicata (Costantini et al. 2006), which represented the actual forest cover extension, and also included the area subject to natural forest expansion. The five year difference between the Forest Map (2006) publication and the change detection analysis (2010) resulted in the application of a 100 m buffer to identify any pixel with $\Delta\text{NDVI} > +1.5\sigma$ that can be considered as a possible natural forest expansion during this period. The buffer value was chosen on the basis of the spatial characteristics of forest cover that in many cases is highly fragmented with high perimetral values. Finally, the pixels characterized by $\Delta\text{NDVI} > +1.5\sigma$ within the conifer afforestation and plantation forest types were excluded.

Eighty points were selected by a random stratified sampling approach to validate the proposed methodology. The study area was divided into a 25 x 25 km grid (Fig. 4), with

Tab. 2 - Threshold value accuracy assessments. The producer's accuracy refers to the probability of land-cover classification accuracy. (a): Error of natural expansion of forest classified as others; (K): Cohen's Kappa inter-rater agreement coefficient.

Threshold	Producer's Accuracy	Overall Accuracy	Error ^(a) (%)	Overall K
1·σ	0.88	0.83	13.9	0.23
1.5·σ	0.95	0.94	5.1	0.61
2·σ	0.8	0.77	25.2	0.1

four validation points for each cell randomly chosen among the pixels with $\Delta\text{NDVI} > +1.5\sigma$. The random stratified sampling approach ensured good cover over the regional forest areas, and at the same time randomness of points within each grid cell. Each of the 80 points was ground-checked using GPS to validate the presence of natural forest expansion processes.

Results and Discussion

Natural expansion of forests in the Basilicata region study area was examined by the NDVI differencing approach. According to the purpose of the study, only positive variations of NDVI were considered, with the exclusion of negative changes that were considered negligible. In particular negative changes of NDVI, in terms of loss of forest cover, are mainly represented by forest fires accounting yearly for the 0.1% of total forest cover. Moreover, no deforestation processes are reported according to the Regional law which prevents any land use change on forest ecosystems.

The main parameters of the distribution of ΔNDVI image values are reported in Tab. 1. Mean, mode, and median values were close to zero, as well as skewness, indicating a good distribution symmetry (Fig. 5). The goodness-of-fit of the ΔNDVI image values to a normal distribution was also verified by Kolmogorov-Smirnov non-parametric test. Although the KS test indicates a departures from normal distribution ($p > 0.01$), it is still within the range normally accepted in the literature in order to be considered a normal or near-normal distribution.

Threshold detection was carried out through visual analysis of Landsat TM RGB band compositions and ratios and their comparison across the two year considered (Fig. 6). The area of actual natural forest expansion was also better identified by comparing the digital aerial orthophotos (1994-2008 - Fig. 7). The best accuracy in the detection of natural forest expansion was obtained using a threshold of $1.5 \cdot \sigma$ with $\Delta\text{NDVI} = \mu + 1.5 \cdot \sigma$, as reported in Tab. 2.

This result is consistent with that obtained by similar studies using the same methodology. In particular, the above test and trial approach was applied to detect vegetation changes due to human-induced rapid land use alterations in tropical forest ecosystems as reported by Miller et al. (1978) in Thailand, Phuaa et al. (2008) in Malaysia, and Sinch (1986) in India based on Landsat imagery. In addition, Hayes & Sader (2001) in Guatemala, and Petit et al. (2001) and Braimoh & Vlek (2004) in Africa applied the same technique to detect forest clearing and vegetation regrowth.

The trial and test approach was also used in forest management for clear-cuts and regrowth detection in temperate forest ecosys-

Fig. 6 - Comparison of Landsat TM RGB band compositions (TM4-3-2) of two test areas for the year 1984 (a, b) and 2010 (c, d).

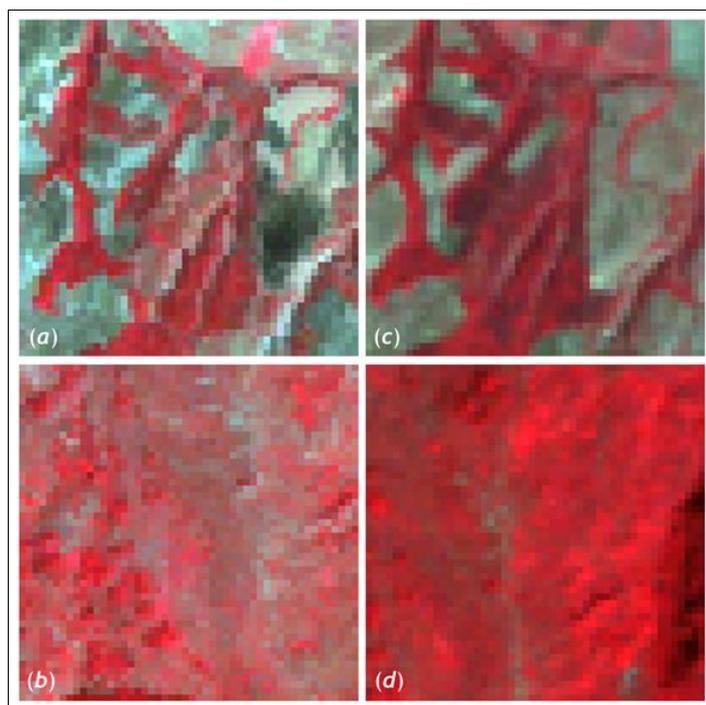
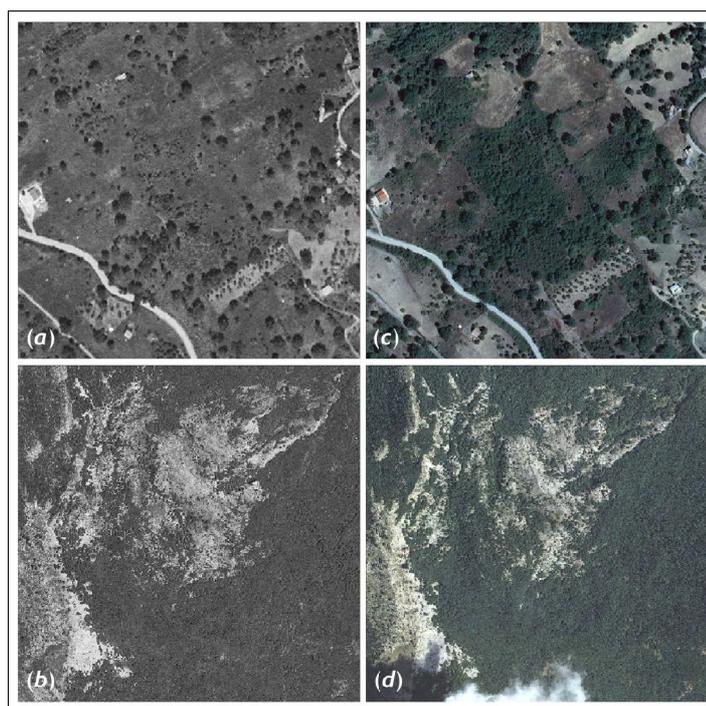


Fig. 7 - Comparison of digital aerial orthophotos for two test areas based on images from *Volo Italia 1994* (a, b) and *TerraItaly 2008* (c, d).



tems (Banner & Lynham 1981, Hayes & Sader 2001), and for the identification of selective logging (Win et al. 2012). Sepehry & Liu (2006) used symmetric threshold identification by testing threshold values ranging between $1 \cdot \sigma$ and $2 \cdot \sigma$ to monitor land cover change induced by flood, while Sarp (2012) used $1 \cdot \sigma$ threshold to detect vegetation change and main environmental effects of surface mining activities. Jomaa & Bou Keir (2003) reported anthropogenic land use change detection with the trial and test approach, and determined $2 \cdot \sigma$ was the optimal threshold value for two sites in Lebanon. Yacouba et al. (2009) applied $1 \cdot \sigma$ in Yunnan province (China), while Podeh et al. (2009) and Coban et al. (2010) ascertained $2 \cdot \sigma$ was the optimal value for land use change detection over complex vegetation cover. Finally, Pu et al. (2008) tested different threshold values ranging between $1 \cdot \sigma$ and $2 \cdot \sigma$ to determine real changes in sparse vegetation cover composed of shrubby trees colonizing riparian areas in Nevada. In some cases, the threshold identification was obtained from extreme percentile values of the NDVI difference image for evaluation of vegetation changes in drainage basin forests (Ferrarini et al. 2000, Mandrone et al. 2006), or forests in hydrogeologically unstable areas (Del Barba et al. 2006). Other more complex threshold identification methods have been recently proposed (Metternicht 1999, Bruzzone & Fernandez Prieto 2000a, 2000b, Hodgson et al. 2004, Im et al. 2007, Boone et al. 2007, Almutairi & Warner 2010). Nevertheless, due to its simplicity, the image thresholding method based on standard deviation remains the most commonly used method for the detection in vegetation changes.

Based on the results obtained in this analysis, the $\Delta\text{NDVI} > +1.5 \cdot \sigma$ image was extracted and clipped using the Basilicata Region Forest Map mask with a 100 m buffer as specified in the Material and Methods. The resulting image (Fig. 8) shows the areas of likely natural forest expansion used for the identification of the ground validation points based on the random stratified sampling.

Among the 80 ground validation points selected, only six were not accessible due to topographic constraints. As for the remaining 74 points (Tab. 3), the vegetation changes detected were the result of natural forest expansion processes in 91.9% of the cases (68 points), while in 8.1% (6 points) the proposed method failed to detect natural forest expansion. In two cases, the failure was due to the coppice revegetation after harvesting; in a single case, the vegetation change was determined by conifer afforestation in a degraded broadleaf forest. Other sources of error were related to revegetation after a fire event in Mediterranean Macchia (one point), and incorrect attribution of land

Tab. 3 - Ground-truth validation results.

Cover type	Samples (n)	Accuracy (%)	Characteristics
Natural afforestation	68	91.8	-
Others	6	8.1	Forest harvesting (2), conifer afforestation in broadleaf forest (1), revegetation after forest fire (1), other land covers (2)
Total	74	100	-

Tab. 4 - Comparisons between natural afforestation from NDVI differencing methodology and inventory forest cover increments.

Method	Forest cover increment ha (%)	Forest cover annual increment (ha yr ⁻¹)
NDVI differencing (1984 - 2010)	70154 (19.7)	2698
National Forest Inventories (1984 - 2005)	62126 (17.4)	2958
Difference between methods	8028 (2.3)	-

cover classes (an olive grove in one case). One last source of error was related to a strong NDVI increase within a forest stand, likely the result of a significant increase in forest density, or to enhanced environmental conditions for growth in recent years. In this specific case, the source of error could be related to quality differences between the two images analyzed. Overall, the accuracy of the ground-based validation reported in Tab. 3 is consistent with that reported in other studies using the NDVI differencing tech-

nique (Hayes & Sader 2001, Podeh et al. 2009, Anderson 2008, Pu et al. 2008, Wang & Xu 2010).

Our results indicated that in the period 1984-2010 approximately 70 154 ha were affected by natural forest expansion processes in the study area, with an annual forest cover increase of 2698 ha (Tab. 4). Given the actual forest cover extension (355 367 ha) as provided by the Basilicata Region Forest Map (Costantini et al. 2006), the increase of forest cover between 1984 and

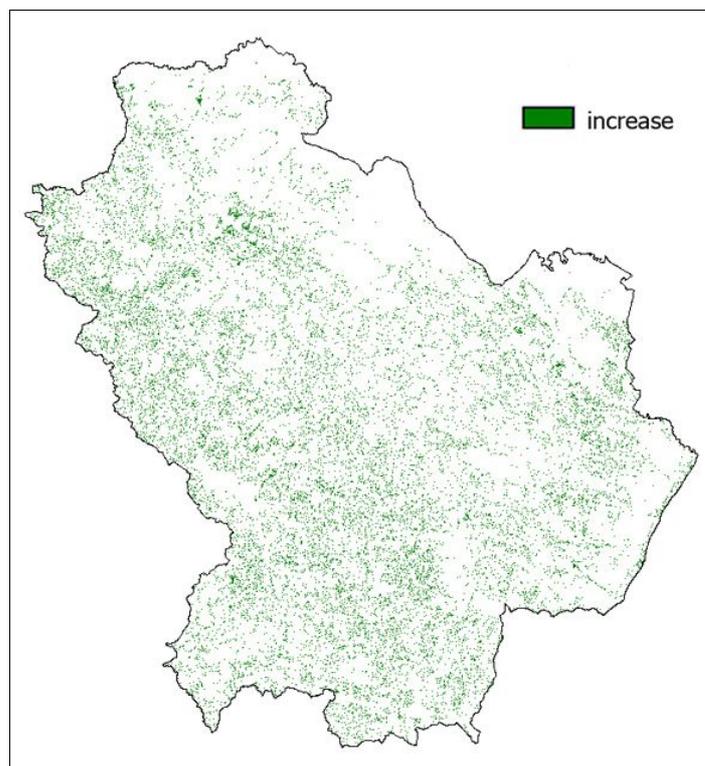
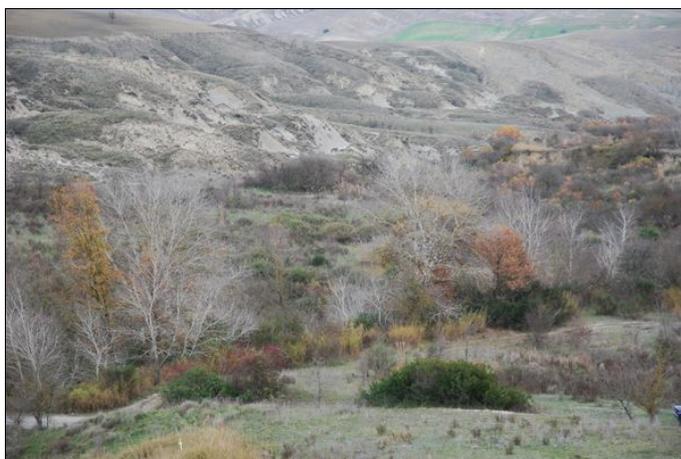


Fig. 8 - Probable areas of natural forest expansion in the Basilicata region.

Fig. 9 - Natural forest expansion on abandoned cattle-raising area for Beech forest in the Pollino National Park area.



Fig. 10 - Natural forest expansion of the Mediterranean macchia vegetation type in a drought-prone Calanchi area.



to natural forest expansion. Although 48.9% of the increment was due to deciduous oak forests, a consistent part of it occurred in the meso/thermophiles component of this vegetation type. Furthermore, natural forest expansion occurred for the Mediterranean macchia (11.4%) and garrigue (2%). Beech forests (Fig. 9) and Mediterranean macchia (Fig. 10) in drought-prone Calanchi areas also showed a natural forest expansion in cattle-raising abandoned areas generally covered by shrubbery followed by tree restoration. The above results highlight the natural forest expansion occurring on degraded and drought-prone soils of the Mediterranean areas, a regions considered highly vulnerable to desertification. Due to their intrinsic ecological interest, these areas require further study to examine their vegetation potential and dynamics.

Conclusions

The present study demonstrated how satellite-based detection of vegetation change can provide reliable results in the assessment of natural expansion of forests and of forest cover dynamics. The validation of the method used by visual analysis of aerial orthophotos and comparison with ground-level sampling control points allowed a reliable identification of vegetation changes due to natural forest expansion. Finally, the relative simplicity of the methodology and the availability of time series Landsat TM images at low cost favors the application of the approach described to large scale forest inventories. Our results also emphasized the current natural forest expansion occurring in drought-prone Mediterranean areas, always considered as highly vulnerable to desertification, and therefore difficult to be managed for purposes of environmental protection. The information on forest cover dynamics provided in this study can be considered a useful starting point to further analyze spatial and temporal patterns of vegetation changes in degraded areas.

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2010 was approximately 19.7%. Similar figures may be obtained by comparing forest cover extensions from two consecutive National Forest Inventories carried out in Basilicata over the same period (294 300 ha - IFN 1988; 356 426 ha - INFC 2005). The difference between the two above inventories (62 126 ha of increase in forest cover) was very close to the value reported in the present study (70 154 ha). The difference between the two latter estimates (8028 ha) may be explained by a further forest expansion occurred over the five years intervening between the last National inventory (2005) and the 2010 Landsat TM image used as reference in this analysis.

Values of natural forest expansion obtained in this study are slightly higher than those reported by the IUTI (*Inventario dell'Uso delle Terre*) land use inventory for the Basilicata region in the period 1990-2008 (Corona et al. 2012). However, this discrepancy may be explained. First, the overall forest cover extension reported by the IUTI inventory for the study area was lower (331 667 ha, approximately 0.7%) than that provided by the Basilicata Region Forest Map (-6.7%), since the latter was based on continuous field recognition rather than on sampling points as for the former. Further-

more, the different time span analyzed (1990-2008 vs. 1984-2010) may accidentally include different trends in main drivers of natural forest expansion, such as forest growth patterns or land abandonment in marginal areas.

Our results also emphasizes some important characteristics of the natural forest expansion in the study area. Indeed, it occurs in areas characterized by severe climate, morphological, and pedological conditions. In 11.4% of the cases analyzed, forest expansion has been detected in areas affected by extreme temperatures (annual mean temperature between 16-18 °C), and in 28.3% of cases in areas with average annual rainfall of 500-700 mm. According to the Bagnouls & Gaussen (1953) index (BGI), 20.8% of natural forest expansion occurred in dry areas with BGI 50-130. These results were also cross-classified with the ESA (Environmental Sensitive Areas) map (Basso et al. 1999, Kosmas et al. 2000, Ferrara et al. 2005, 2010), which presented trends in natural forest expansion. According to the ESA classification, natural forest expansion occurred within the vulnerability class F (Fragile - 29.7%) and class C (Critical - 3.2%).

Our results also support some interesting consideration on the vegetation types subject

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