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## Saproxylic beetles in non-intervention and coppice-with-standards restoration management in Meerdaal forest (Belgium): an exploratory analysis

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For many centuries, coppice-with-standards management was applied in the mixed oak stands of Meerdaal forest (Belgium). Over the last century, these stands were gradually converted to high forest. On an area of 20 ha, the coppice-with-standards management is being restored, with specific adaptations for biodiversity (conservation of dead wood and veteran trees). A survey of saproxylic beetles was performed at 8 locations in the forest, including one site within the coppice-with-standards restoration. This survey not only allowed an evaluation of the saproxylic beetle richness of the forest complex, but also made exploratory observations on the effect of this type of management, as compared to non-intervention, on species richness and composition of saproxylic beetles. The results show that the overall species richness in the forest complex was quite high and comparable to forest reserves in Germany. Both coppice-with-standards and high forest options appear to be equally species-rich, but consist of different communities, both containing specific, rare and notable species, with more thermophilous and light-demanding species in the coppice-with-standards plot. Based on these observations we suggest that a diversified management approach may be the most suitable to conserve and enhance diverse saproxylic beetle communities in formerly intensively managed semi-natural woodlands. This could include areas of active conservation management aimed at producing open-canopy stands with considerable amounts of sun-exposed deadwood, combined with areas of non-intervention in a matrix of multifunctional forests, where conservation of dead wood and veteran trees is fully incorporated in the management.

**Keywords:** Coppice-with-standards, Saproxylic Beetles, Insect Biodiversity, Type of Management, Active Conservation Management

### Introduction

There has been some debate as to whether non-intervention is always the best management option for conserving biodiversity, especially in former traditionally managed landscapes (e.g., Schnitzler 2014 versus Agnoletti 2014). For saproxylic beetles, the amount and variability of dead

wood, especially coarse woody debris (CWD) is considered to be the most important factor determining species richness (Martikainen et al. 2000, Brunet & Isacsson 2009, Müller & Bütler 2010, Stokland et al. 2012, Lassauce et al. 2011). Strict reserves (with non-intervention management, also called a “hands-off approach”) contain or

build up greater quantities and a higher diversity of dead wood than regularly managed forests (e.g., Vandekerkhove et al. 2009a) and correspondingly richer communities of saproxylic beetles (Müller et al. 2007, Paillet et al. 2010, Lachat et al. 2013). Non-intervention is therefore strongly advocated for the conservation of saproxylic communities, including saproxylic beetles (Martikainen et al. 2000, Parviainen et al. 2000, Müller et al. 2007, Bütler & Lachat 2009, Bouget et al. 2014). Such studies, however, often focus on Boreal and Alpine forests, or on beech-dominated forest types.

For forests with a long management history, especially those where coppice management and forest grazing have resulted in open-canopy woodlands, often in lowland and thermophilic oak forests, the “hands-off” strategy is challenged. Hédl et al. (2010) came to the conclusion that spontaneous succession in thermophilic oak forests led to more common mesic forest types, resulting in the loss of many species

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that are specific to this forest type. They suggested reintroducing active coppice management or forest grazing to reverse the succession and restore the specific plant communities of traditionally managed oak forests. Other studies also indicate the loss of vascular plant species richness when coppice-with-standards forests are converted to high forest (Van Calster et al. 2008, Baeten et al. 2009). The positive effects of traditional coppicing on breeding birds and butterflies are often described (Buckley 1992, Fuller & Warren 1996) but are also debated (Hamblen & Speight 1995). In the case of saproxylic beetles, it is argued that non-intervention is not always the best choice, especially in these old oak-dominated forests, as they contain many thermophilous species adapted to the open-grown status of their habitat.

Open canopy conditions – regularly occurring in coppice-with-standards forests – result in specific abiotic conditions, with higher summer temperatures in sun-exposed areas that may be essential for the survival of certain thermophilous species. In addition to the dead-wood resource, temperature is assumed to be a crucial driver of the diversity of saproxylic beetles (Lachat et al. 2012, Vodka & Cízek 2013, Müller et al. 2014). Higher summer temperatures may even compensate for lower dead wood amounts (Köhler 2014, Müller et al. 2014).

Gärdenfors & Baranowski (1992) concluded that 70% of the Swedish red-listed saproxylic beetles living on oaks preferred sun-exposed habitat and only 16% preferred dense stands (while this was the opposite for beech-related species). Several studies (Franc & Götmark 2008, Widerberg et al. 2012, Gough et al. 2014) concluded that a partial opening up of old oak stands might increase the species richness of saproxylic beetles. Some authors even stated that natural succession might be a threat to the specific species community and that a “hands-off” conservation strategy was the “worst solution” and “especially damaging” when applied to traditionally managed woodlands, and should therefore be avoided (Sebek et al. 2013, Miklín & Cízek 2014, Horak et al. 2014). These

studies mainly focus on former wood pastures, still containing old open-grown veteran trees being gradually overgrown by understorey trees following the cessation of mowing and grazing (Ranius & Jansson 2000, Franc & Götmark 2008, Gough et al. 2014, Horak et al. 2014) or interplanted with spruces (Widerberg et al. 2012). However, few studies are known which actually compare the effect of management alternatives (Lassauce et al. 2012, Sebek et al. 2015) or natural disturbances (Bouget 2005) on the species richness of saproxylic beetles in former coppice-with-standards oak woodlands. Some authors advocated non-intervention (Hamblen & Speight 1995, Lassauce et al. 2012), while others – often focusing on thermophilous species – called for reactivation of neglected coppice (Barbalat & Gétaz 1999, Sebek et al. 2015).

In 2010-2011, a survey of saproxylic beetles was performed in the Meerdaal forest (Belgium), a relatively large ancient woodland site consisting mainly of uneven aged mixed oak stands (*Quercus robur* and *Quercus petraea*) that originated from a former coppice-with-standards regime. This survey was part of a larger project to evaluate whether forests in Northern Belgium, with their long history of fragmentation and intensive management (Verhulst 1995, Tallier 2004, Hermy & Verheyen 2007, Vandekerckhove et al. 2009b, De Keersmaecker et al. 2015), still contained a high diversity of saproxylic beetles. In these forests very few veteran trees and virtually no dead wood were present (Vandekerckhove et al. 2011). Only in the past 30 to 40 years has the ecological value of dead wood and veteran trees gained attention in forest management and progressively been reinstated (Vandekerckhove et al. 2011). This management history and lack of continuity in dead wood may have resulted in an impoverished fauna, containing only good dispersers with limited substrate requirements, but lacking the more demanding and dispersal-limited species. However, no extensive datasets were available in Belgium to underpin this hypothesis, so specific surveys were set up to analyze the potential of forests in Northern Belgium for the conservation of saproxylic species in woodland

sites that were considered to be of high potential for saproxylic beetles, including Meerdaal forest (Vandekerckhove et al. 2011). As the Meerdaal forest survey encompassed both unmanaged plots and the coppice-with-standard restoration, it also allowed for some preliminary observations on the effect of this management intervention on the saproxylic beetle community.

## Material and methods

Meerdaal forest is an ancient woodland site, located 30 km east of Brussels, covering an area of approximately 1300 ha. The prevailing sub-Atlantic climate of this region has a total annual precipitation of ca. 821 mm (uniformly distributed over the year), and minimum and maximum average monthly temperatures of 2.5 and 17.2°C, respectively, with a yearly average temperature of 9.7 °C. The forest is located on a slightly undulating plateau, with an elevation between 55 and 105 m a.s.l. Predominant soil types in Meerdaal forest are Luvisols (in the south) and Podzoluvisols (in the north), which developed in Pleistocene aeolian deposits of loamy loess on top of tertiary sandy formations (Baeyens et al. 1957). Main vegetation types are *Stellario-Carpinetum* and *Milio-Fagetum* (*sensu* Noirefals 1984) on Luvisols. On Podzoluvisols, the predominant vegetation type is *Fago-Quercetum* (*sensu* Noirefals 1984).

The forest mainly consists of uneven-aged oak stands, complemented with beech (*Fagus sylvatica*) and conifer stands (mainly *Pinus sylvestris* – out of the scope of this study). The beech stands were mostly planted in the period 1850-1900, and are managed as high forest with regular thinning. For centuries, the oak stands were managed as coppice-with-standards but were gradually converted to high forest over the last 100 years (Vandekerckhove et al. 2016).

In 2010-2011 an intensive inventory of saproxylic beetles was performed in this forest complex. The sampling sites were specifically chosen at locations with a high “potential” for saproxylic beetles, *i.e.*, on sites with relatively high amounts of dead wood and old and decaying trees. Also differences in light conditions were considered

Tab. 1 - Overview of the sampling sites in Meerdaal forest.

No	Description	Sampling year	Basal area (m <sup>2</sup> ha <sup>-1</sup> )	Dead wood volume (m <sup>3</sup> ha <sup>-1</sup> )
1	Mixed oak-hornbeam forest; dominant tree layer = oaks (100-250 years old) - strict reserve since 1995	2011	34.5	29.1
2	Mixed oak-hornbeam forest; dominant tree layer = oaks (100-250 years old) - strict reserve since 1995	2011	35.5	26.6
3	Coppice-with-standards restoration - standard trees of oak (100-250 years old) - restoration cut in winter 2010-2011	2011	12.9	21.0
4	Even-aged beech stand (145 years old) - strict reserve since 1995	2011	32.5	33.5
5	Even-aged beech stand (110 years old) - strict reserve since 1995	2010	24.5	18.5
6	Mixed oak-hornbeam forest; dominant tree layer = oaks (100-250 years old) - strict reserve since 1995	2010	31.4	36.9
7	Beech tree alley (ca. 130 years old), outside but connected to the forest	2010	-	-
8	Mixed oak-hornbeam forest; dominant tree layer = oaks (100-250 years old) - strict reserve since 1995	2010	34.7	28.1

red, covering both closed canopy non-intervention and open, sun-exposed sites.

In 8 localities (4 each year) a standardized survey was performed. It comprised of 6 non-intervention sites (4 oak and 2 beech stands), an old beech tree alley and one site in a coppice-with-standards restoration project (Tab. 1).

The non-intervention sites were located in strict forest reserves, 20-50 ha in size, that were established in 1995. The two beech non-intervention sites were even-aged, monospecific beech stands of 110 and 145 years old, previously managed as high forest and regularly thinned. The oak stands consisted of mixed high forest originating from coppice-with-standards, with a relatively high number of standard trees. Since the 1950s, when coppicing was abolished, the oak stands have been managed as mixed, uneven-aged high forest with regular thinning and group felling. The stands are dominated by an uneven-aged layer of the former standard oak trees, between 100 and 250 years old, combined with a dense sub-canopy mainly of maple (*Acer pseudoplatanus*) and hazel (*Corylus avellana*). The stands in strict reserves have now been left unmanaged for over 15 years.

One sampling site was located in a mixed oak stand of 20 ha that was declared a “managed forest reserve” in 1998, in order to restore the coppice-with-standards management. This restoration was subdivided into eight subplots of 2-3 ha, of which one subplot was intensively thinned every second year in order to re-create the forest structure (species composition, stem density, diameter distribution and basal area) of 100-200 years before. Modifications to the traditional management are that veteran and dead trees present in the stand were all conserved, as was a share of the felling residues. Sampling of this site was performed within the coupe of winter 2010-2011, in which the basal area was reduced from ca. 25 to 13 m<sup>2</sup> ha<sup>-1</sup>. It is located adjacent to the coupe of 2008-2009.

The beech tree alley consisted of 120-150 year old beech trees, and is located outside but connected to the forest. It contains several moribund trees, with dead branches in the canopy and on the ground. Data on basal area and local dead wood volumes in Tab. 1 are derived and updated from dendrometric inventory data compiled in the forest reserves monitoring program (Vandekerckhove et al. 2003) and the management plan of the Meerdaal forest.

Beetle sampling was done from April to November and involved both active and passive sampling. The latter consisted of a window trap and a glue-ring, both attached to one dead or dying standing tree per site. Active sampling aimed to be comparable at each site and involved 3 sampling sessions of ½ hour per site (in May, June and July) by two people performing sweep netting and beating tray catches on wood, fungi and shrubs and the collection

and analysis of 5 wood-mould sieve samples (of comparable size) per site.

All *Coleoptera* were identified to the species level. They were subdivided into forest-related species versus eurytope and non-forest species. The forest-related species were divided between saproxylic and non-saproxylic species according to Köhler (2000). The saproxylic species were subsequently allocated to six substrate classes (Köhler 2000). These classes distinguish species living in the early decay stage of dead wood (wood), in highly decayed wood mould (mould), in decaying bird and insect nests (nest), in wood-decaying fungi (fungi), cambium feeders and predators living in and under bark (bark) and species primarily living or depending on sap run in living trees (sap run). The conservation status of the saproxylic species was assessed based on the Red Lists for Germany (Geiser 1998) and for Europe (Nieto & Alexander 2010), as no Red List for Belgium exists. Indicator species for old-growth continuity were derived from the list developed by Müller et al. (2005).

The results of the survey allowed for an overall evaluation of the saproxylic beetle richness of the forest complex, but also to compare the separate sites, which were all sampled in the same way.

Although the sampling setup was not originally designed to this purpose – the original aim was to evaluate the current overall saproxylic beetle richness of the forest complex –, this survey in Meerdaal forest also allowed to perform some exploratory observations on the saproxylic beetle richness of two conservation alternatives, non-intervention and restoration of coppice-with-standards, in a forest that was previously much more open and intensively managed for wood production.

Comparisons were made in a descriptive way, comparing total species richness, red-listed species richness, and the number, abundance and characteristics of the species that were exclusively found in one site.

Comparability of the sites was further analyzed using ordination. Separate analyses were made for the complete dataset of saproxylic beetles using all sampling techniques (active and passive) and including only the standardized passive catch methods (glue ring and window trap). Presence/absence data was used with the Jaccard method as distance measure. A permutational multivariate analysis of variance (PERMANOVA) was performed using the Anderson method (Anderson 2001) to evaluate the effect of the following factors: (a) tree species; and (b) management (non-intervention, tree alley and coppice-with-standards). Effects of both variables were tested in separate models. As the sampling was spread over two years, we also checked for sampling year bias by running the PERMANOVA on the subset of non-intervention sites (2 oak, 1 beech per sampling year), and adding the factor “year” to the full model.

The Anderson method uses a permutation test with pseudo-F ratios to fit linear models to distance matrices (Anderson 2001). PERMANOVA models were performed with 1000 permutations in R 3.2.2 (R Core Team 2015) using the library “Vegan” (Oksanen et al. 2013). The graphs were made with the Capmod function in Vegan.

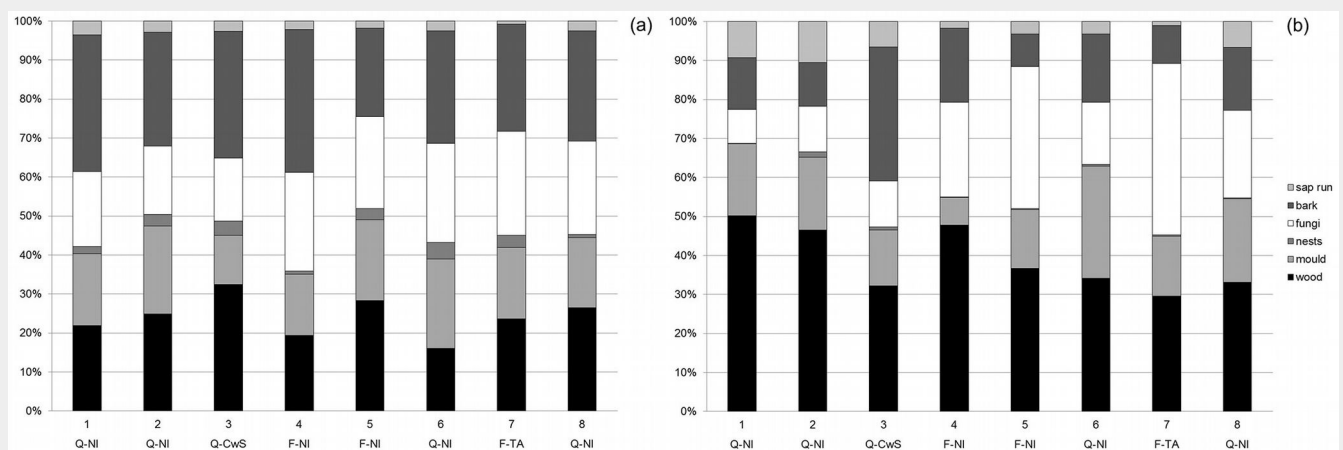
## Results

In total, 22 390 adult beetles of 781 species were captured, of which 10 583 individuals belonged to 294 species that are considered saproxylic according to Köhler (2000). Seventy-six of these saproxylic beetles are listed in the German Red List (Geiser 1998), of which four are in the “critically endangered” category (*Aeletes atomarius*, *Triplax rufipes*, *Plectrophloeus rhenanus* and *Anaspis garneysi*), 21 are “endangered” and 51 are “vulnerable”. One species living in brown rot wood-mould, *Cerophytum elateroides*, appears in the European Red List (Nieto & Alexander 2010) in the category “vulnerable” (but “endangered” on the German Red List). Another species, *Aeletes atomarius*, is an indicator of dead wood continuity, the so-called “Urwald-relict-species” (Müller et al. 2005).

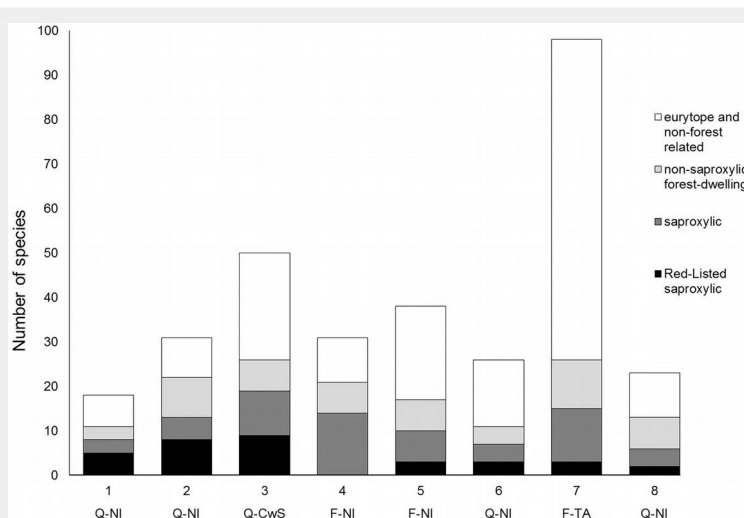
An overall descriptive comparison for the eight sampling sites is given in Tab. 2. When comparing the total species richness (all beetle species, including ubiquitous) in the eight sampling sites, similar figures of 230-280 species per plot were detected, except for site 7 (the tree alley) where 379 species were found. However, if we exclude non-forest-dwelling and non-saproxylic species, then the differences are much less pronounced, with numbers varying between 106 and 137. Beech and oak-dominated sites had comparable species richness, and open canopy sites (3 and 7) were not particularly richer than the closed-canopy sites (Tab. 2). Dominant tree species and site openness did not appear to determine the overall species richness of the different sampling sites, nor the number of red-listed species. Larger differences in abundance can be explained by one or two species being much more abundant in one of the sites. For the coppice-with-standards site, the ubiquitous carabid beetles *Bradycellus verbasci* and *Bradycellus harpalinus* are much more abundant than in other plots and make up more than half of all captured specimen at this site. Also, when we compare the partitioning of species over different substrate types, we do not see any clear pattern related to tree species or sun-exposure. (Fig. 1). Based on the literature (Köhler 2000, 2011, Bouget et al. 2014), it might be expected that open sites (i.e., coppice-with-standards restoration and tree alley) as compared to the closed-canopy non-intervention plots, would contain relatively more wood and bark related species preferring fresh sun-exposed substrate and less fungi-related species, the fungi being associated with more shaded and moist conditions. With regards to spe-

**Tab. 2** - Results for the 8 sampling sites, both on abundance (#indiv) and species richness (#spp), for all beetles, saproxylic beetles only, red-listed species (RL1: critically endangered; RL2: endangered; RL3: vulnerable) and for selected and subfamilies. Tree species and management codes: (Q-NI): oak stands with non-intervention; (Q-CwS): oak stands with coppice-with-standards management; (F-NI): beech stands with non-intervention; (F-TA): beech tree alley in open countryside.

Group	Tree species and management	Sampling site no.								Total
		1	2	3	4	5	6	7	8	
All Coleoptera	# indiv	2267	1899	3220	2622	2739	3160	4678	1769	22354
	# spp	230	281	267	269	257	276	379	256	781
All Saproxylic	# indiv	1098	1310	859	2009	1576	1468	1456	807	10583
	# spp	114	137	111	134	106	118	131	117	294
RL1	# indiv	0	0	28	149	64	1	5	19	266
	# spp	0	0	1	1	2	1	3	2	4
RL2	# indiv	28	31	15	14	19	47	27	15	224
	# spp	5	8	8	2	4	6	3	4	21
RL3	# indiv	74	206	110	90	340	207	208	83	1318
	# spp	17	29	11	18	22	18	25	19	51
Staphylinidae	# indiv	71	212	93	197	105	127	60	73	938
	# spp	26	24	17	26	17	21	21	16	52
Buprestidae	# indiv	4	0	216	0	0	0	0	0	220
	# spp	2	0	4	0	0	0	0	0	4
Cerambycidae	# indiv	19	8	35	9	18	3	14	22	128
	# spp	6	8	10	7	8	1	4	4	22
Scolytinae	# indiv	542	546	159	898	40	511	129	159	2984
	# spp	9	10	12	12	4	9	11	8	20
Latridiidae+Mycetophagidae+Ciidae	# indiv	80	98	48	179	396	203	556	120	1680
	# spp	9	11	6	16	14	18	19	13	26



**Fig. 1** - Relative share of species (a) and specimens (b) including both active and passive catches, allocated to different substrate classes (according to Köhler 2000) for the 8 sampling sites of Meerdaal forest. Tree species and management codes: (Q-NI): oak stands with non-intervention; (Q-CwS): oak stands with coppice-with-standards management; (F-NI): beech stands with non-intervention; (F-TA): beech tree alley in open countryside.



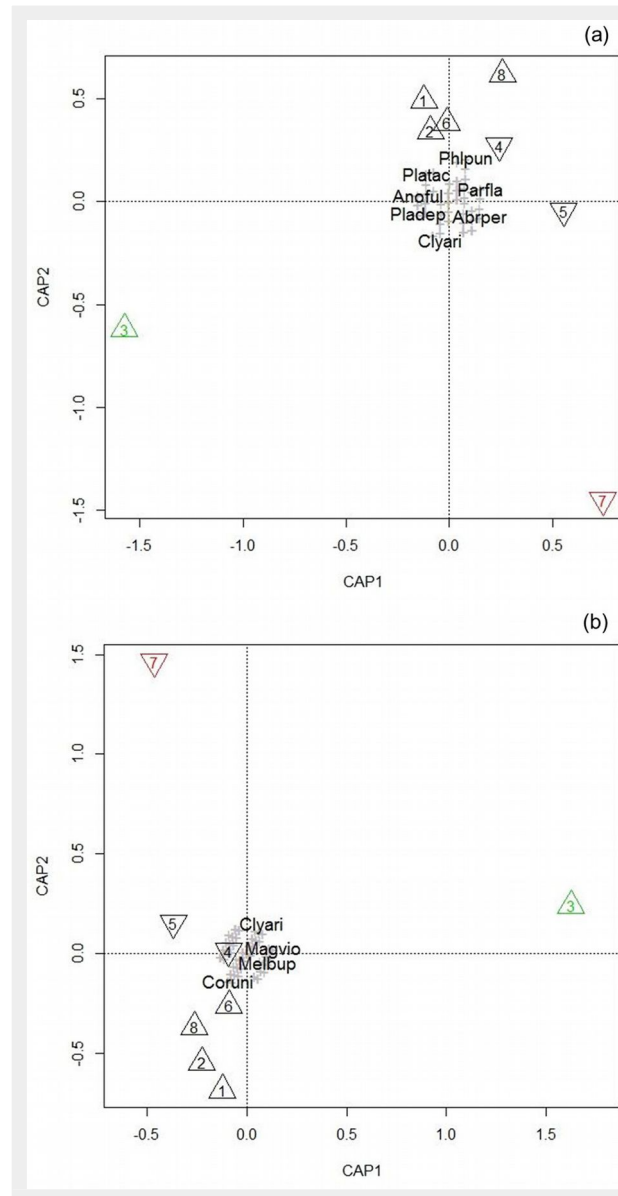
**Fig. 2** - Number of exclusive *Coleoptera* species for every plot, subdivided in red-listed saproxylic species, non-red-listed saproxylic species, non-saproxylic forest-dwelling species and eurytope and non-forest beetle species (active+ passive catches). Tree species and management codes: (Q-NI): oak stands with non-intervention; (Q-CwS): oak stands with coppice-with-standards management; (F-NI): beech stands with non-intervention; (F-TA): beech tree alley in open countryside.

cies richness (Fig. 1a), the coppice-with-standards restoration appears to meet these expectations, but the tree alley does not. As for relative abundance (Fig. 1b), the coppice-with-standards restoration clearly contained relatively more bark-related species, but this can be attributed to the relative abundance of one species (*Agrilus sulcicollis*) that made up more than 10% of all captured saproxylic beetles in this plot. The tree alley showed a different pattern with more fungi-related individuals, also explained by the high abundance of two species (*Octotemnus glabriculus* and *Cis castaneus*).

However, differences did emerge when comparing the species richness and number of individuals for some selected families in Tab. 2. For *Staphylinidae*, the most species-rich family in this survey, including species of diverse habitat groups and both sun-exposed and shaded conditions, no obvious differences were apparent. Likewise, for *Scolytinae*, more exclusively linked to bark substrates, no specific link between diversity and tree species or sun-exposure was observed. However, for the families with a pronounced preference for sun-exposed wood, jewel beetles (*Buprestidae*) and longhorn beetles (*Cerambycidae*), the coppice-with-standards site appears to be richer than the other sites, both in species diversity and number of individuals. For jewel beetles in particular, the difference was striking. Although also sun-exposed, the beech alley did not show a similar pattern: species richness for the above-mentioned families was comparable to the non-intervention sites.

For families consisting mainly of small species living in saproxylic fungi and wood-mould, often in shaded conditions, like *Latriidae*, *Mycetophagidae* and *Ciidae*, an inverse pattern could be observed, though less pronounced. For these families, the coppice-with-standards site showed the lowest figures, both in species richness and abundance.

We also checked the number of exclusive species for each site (Fig. 2). If comparing all beetle species, including non-forest and non-saproxylic species, then the beech alley unmistakably contained the highest number of exclusive species, as this site also captured many eurytope species from the open countryside. When considering only forest-dwelling and saproxylic species, then the coppice-with-standards site had the highest number of exclusive species for forest-related species, as well as saproxylic and red-listed saproxylic species. The species exclusively found in the coppice-with-standards plot included the longhorn beetles *Xylotrechus rusticus*, *Xylotrechus antilope*, *Plagionotus detritus*, *Plagionotus arcuatus* and *Phymadotes alni*. Other species were *Cerophytum elateroides* and the jewel beetles *Agrilus biguttatus* and *Agrilus laticornis*. All but one of the 92 specimens of *Agrilus angustulus* and all but 3 of 90 the specimens of *Agrilus sulcicollis* were caught



**Fig. 3** - Capmod graph of the PERMANOVA model for both active and passive catches (a) and including only the results of the passive catches (b). Triangle: oak; reversed triangle: beech; black: unmanaged; green: coppice-with-standards; red: tree alley. Species are represented by a grey cross or abbreviated name. (Abrper): *Abraeus perpusillus*; (Anoful): *Anobium fulvicorne*; (Clyari): *Clytus arietis*; (Coruni): *Corticteus unicolor*; (Magvio): *Magdalis violacea*; (Melbup): *Melasis buprestoides*; (Parfla): *Paromalus flavicornis*; (Phlpun): *Phloeonomus punctipennis*; (Pladep): *Placusa depressa*; (Platac): *Placusa tachyporoides*.

at this site.

The results of the statistical analysis are given in Fig. 3 and Tab. 3. The Capmod-graphs from the PERMANOVA model for both analyses (including and excluding the active catches) gave similar results: the unmanaged oak and beech stands are clustered together, whereas the beech alley and the coppice-with-standards plot were clearly divergent from the rest and each other.

The results of the model (Tab. 3) indicate that the “tree species” factor was not significant, whereas “management”, distinguishing between non-intervention, coppice-with-standards and the tree alley, apparently affects the species composition.

When considering only passive catches, the factor “management” was significant, while for both active and passive catches it was marginally significant.

To interpret the axes of the ordination, the ten most influential species for these axes in the direction of plot 3 and 7 in the Capmod-graphs are listed in Tab. 4. Thermophilous species, especially longhorn beetles – further complemented by *Agrilus* species for the analysis with only passive catches – were prominent in the positioning of the CAP1 axis that strongly differen-

**Tab. 3** - Results from the PERMANOVA model for both active and passive catches (a) and including only the results of the passive catches (b). (\*):  $p < 0.05$ .

Model	Factor	df	F	p
(a) Active + passive catches	Tree species	1	0.99395	0.4945
	Management	2	1.5831	0.07393
(b) Passive catches	Tree species	1	1.1076	0.3986
	Management	2	1.5163	0.03397 *

**Tab. 4** - Lists of the 10 most influential species determining the CAP1 and CAP2-axis of the ordination in the direction of plot 3 and 7 for both active and passive catches (left) and including only the results of the passive catches (right).

Passive + active catch		Passive catch	
CAP1	CAP2	CAP1	CAP2
<i>Mordella aculeata</i>	<i>Ptinella aptera</i>	<i>Malthodes marginatus</i>	<i>Cerylon ferrugineum</i>
<i>Platycerus caraboides</i>	<i>Dasytes plumbeus</i>	<i>Scaphisoma agaricinum</i>	<i>Melanotus villosus</i>
<i>Phymatodes alni</i>	<i>Diplocoelus fagi</i>	<i>Plagionotus detritus</i>	<i>Cryptarcha strigata</i>
<i>Xylotrechus rusticus</i>	<i>Sphindus dubius</i>	<i>Taphrorychus villifrons</i>	<i>Xyloterus domesticus</i>
<i>Xylotrechus antilope</i>	<i>Sulcaxis affinis</i>	<i>Agrilus biguttatus</i>	<i>Euplectus karsteni</i>
<i>Plagionotus detritus</i>	<i>Dorcatoma robusta</i>	<i>Agrilus laticornis</i>	<i>Cetonia aurata</i>
<i>Plagionotus arcuatus</i>	<i>Dorcus parallelipipedus</i>	<i>Aplocnemus impressus</i>	<i>Dacne bipustulata</i>
<i>Crypturgus pusillus</i>	<i>Stereocorynes truncorum</i>	<i>Crypturgus pusillus</i>	<i>Enicmus brevicornis</i>
<i>Taphrorychus villifrons</i>	<i>Magdalis ruficornis</i>	<i>Phymatodes alni</i>	<i>Bitoma crenata</i>
<i>Magdalis violacea</i>	<i>Clytus arietis</i>	<i>Placusa depressa</i>	<i>Clytus arietis</i>

tiated the coppice-with-standards from unmanaged plots, indicating that this axis can be linked to sun-exposure.

The results appeared not to be biased by the year of sampling. We tested the factor “sampling year” in a PERMANOVA on the subset of 6 non-intervention sites (2 oak, 1 beech every sampling year), both with and without tree species as a factor in the model, and with the two subset variants (with and without the active catches). Sampling year was not significant in all 4 cases. Also, when sampling year was added as an additional factor to the PERMANOVA model presented here, this factor was not significant and did not alter the outcome (results not shown).

## Discussion

The overall species diversity of saproxylic beetles in Meerdaal forest was relatively quite high, both in quantity and quality of species. When compared with other forest reserves in Germany and Luxembourg that were surveyed in a similar way (Köhler 2014), Meerdaal forest had a greater than average total species richness of saproxylic beetles: 294 species compared to an average of 256 for 24 sites in Germany and Luxembourg. The number of 75 red-listed species was also in line with the average recorded in these 24 sites.

These figures were higher than expected as the relatively small size and high level of isolation of Meerdaal forest are less favorable for a rich saproxylic beetle community than in most comparison sites. It also has had a long history of very intensive management, with little or no deadwood remaining at the site (Vandekerckhove et al. 2016), and very few veteran trees. In 1913, barely 15 trees over 250 cm in girth occurred in the whole forest complex (Vandekerckhove et al. 2009b). It is uncertain whether saproxylic species managed to survive over past centuries in relic habitats like cavities in old coppice stools (Lassauce et al. 2012), dead branches in the standards or in pollard trees in the surrounding countryside (that have now almost all disappeared), or whether they have managed to recolonize the newly established favorable habitats from longer distances (Jonsson et al. 2005, Vandekerckhove et al. 2011, 2013).

Nevertheless, the presence of only one indicator species of dead wood continuity (“Urwald relic species” according to Müller et al. 2005) indicates that slow dispersers with high habitat requirements were under-represented. This is the case in most European lowland forests, including the German and Luxembourg reserves this site is compared with.

On the other hand, the Meerdaal forest complex now exhibits relatively large amounts of dead wood, especially in the forest reserves. There, the amounts are in the required range of 30-50 m<sup>3</sup> per ha of dead wood for rich saproxylic communities (Müller & Büttler 2010). The forest also contains a high density of very large trees: over 1200 individuals measure over 3 meters in girth, of which more than 500 are oak trees (Vandekerckhove et al. 2009b). Considering only these oaks, the average density (0.35 trees ha<sup>-1</sup>) doubles the threshold of 0.15 trees ha<sup>-1</sup> mentioned by Bergman et al. (2012) required to ensure a rich saproxylic oak fauna. In this context, dispersal-limited species whose habitat requirements are currently met, may still be missing, resulting in a so-called colonization or immigration credit (Hanski 2000, Jackson & Sax 2009).

The inventory in Meerdaal forest, spread over 8 sites and two survey years, was designed to produce a first evaluation of the importance of the forest complex for saproxylic beetles. The survey was not originally set up to compare the effect of different management alternatives on species richness and abundance of saproxylic beetles. To draw full conclusions on this topic, a more equal distribution of the different management types, including regular management, and with more replicates for each management type, is required.

Nevertheless, these results are sufficiently explicit to make preliminary observations on the species richness of saproxylic beetles in the analyzed situations. Species richness in the coppice-with-standards site appeared to be comparable to the non-intervention sites, but species composition in the former was substantially different from both the closed-canopy non-intervention sites and the equally sun-exposed tree al-

ley that was surveyed.

In the applied model, the “management” factor showed to be significant or marginally significant (with and without active catches), while the “tree species” factor was clearly not significant, indicating that the level of sun-exposure was a stronger determinant of species composition in saproxylic beetles than the dominant tree species.

Due to the limitations of the experimental setup we must be cautious with the interpretation of these results. We further tested our observations by analyzing more variants of the model, adding non-saproxylic species and using abundance instead of presence-absence data, again excluding and including the active sampling techniques. For all these variants, the results remained similar: the consistency of the results, irrespective of the different data subsets used, supports our assumption that the “management” factor is a true effect that shapes the beetle species composition. Nevertheless, the restrictions of the dataset only allow us to conclude that the effect is likely. Further research with full replicates should be undertaken to fully prove this effect.

Our observations indicate that the saproxylic beetle community on the coppice-with-standards restoration plot appears to be different from, but not necessarily richer than, that of the non-intervention plots. This observation apparently differs from other studies where saproxylic beetle richness of oak woodlands was found to be higher in sites with more open canopies.

Many of these studies focus on the restoration of abandoned wood pastures in Scandinavia. Ranius & Jansson (2000) compared species richness of cavity-inhabiting saproxylic species on 18 oak sites in Sweden: next to tree size (girth), low canopy cover also increased the frequency of several species. Franc & Götmark (2008) presented the results of a large experiment for saproxylic beetles in which non-intervention was compared to a specific “partial cutting” regime (25% of basal area, retaining both large and dead trees) in paired plots in 22 old oak sites in Southern Sweden. They found on average 35% more species in the partially cut areas, but there was

no difference in the species richness of red-listed species, only in the composition of exclusive species between the managed and the uncut areas. Widerberg et al. (2012) found an increase in species richness and abundance of oak-associated species when old oak trees closed in by spruce plantations were released. Gough et al. (2014) also concluded that clearing of re-growth in areas with veteran oaks was expected to increase species richness of vulnerable and specialist saproxylic beetle species. These results may apply particularly to the management history of the wood pastures, still containing old open-grown veteran trees but now being gradually shaded and suppressed by understorey trees, following the cessation of mowing and grazing management. These wood pastures are known to be very rich in saproxylic species, and are of particular interest to highly demanding species with low dispersal ability and living in wood-mould of sun-exposed, hollow trees (Harding & Rose 1986, Miklín & Cížek 2014, Ranius & Jansson 2000). Natural succession in these situations would ultimately lead to shading and the suppression of the old oaks by late-successional species, which is detrimental to the rare saproxylic community specifically associated with veteran oaks. Restoration of more open canopy, while conserving the veteran oaks and dead wood, will clearly benefit these communities, while the newly arriving species of closed canopy forest will most likely survive in the immediate surroundings. In these circumstances, active conservation management may indeed result in higher species richness. However, Franc & Götmark (2008) warned against too drastic and general application of these interventions. Too strong opening of the canopy may result in increased mortality of the old trees, and may also disadvantage many red-listed species preferring semi-open or closed canopies. They advised a differentiated management approach, including both active management and minimal intervention.

In the original coppice-with-standards forests that were later abandoned or converted to high forest, the situation was quite different: it is unclear what the composition of the saproxylic beetle communities of these intensively managed stands were, and how rich they were. They were also frequently sun-exposed, but contained no veteran trees and hardly any dead wood (Kirby 1992, Vandekerckhove et al. 2011). Cavities were mainly restricted to the old coppice stools that were frequently shaded. It can therefore be assumed that the saproxylic communities in these forests were different from the wood pastures, favoring species with higher dispersal capacity, more adapted to shade and less restricted by continuous sun-exposure, or able to also survive within sun-exposed parts of the upper canopy of closed-canopy forest phases.

Studies that focus on the specific effect of coppice and coppice-with-standards management (or neglect) on saproxylic beetle communities are scarce. A small-scale survey of overmature coppice by Barbalat & Gétaz (1999), comparing freshly cut coupes with a 5-year-old coupe and undisturbed forest, concluded that the recent canopy openings were richer in species than the uncut areas and the 5-year-old coupe. However, they only sampled 4 beetle families, which mainly contained thermophilous species (*Buprestidae*, *Cerambycidae*, *Lucanidae* and *Scarabaeidae*). On the 5 year old coupe, where the canopy had closed again, the positive effect of the intervention was no longer observed. Sebek et al. (2015) also concluded that recently cleared areas in an overmature coppice forest showed a higher total species richness in saproxylic beetles, but not for red-listed species, and indicated that a longer time series would be needed to evaluate the effect in later stages of the coppice cycle. Unfortunately, these studies do not contain any information on dead wood amounts and composition before and after the intervention, making it impossible to evaluate the impact of the clearings on resource availability.

On the other hand, Lassauce et al. (2012) compared still operational coppice-with-standards forests with a mature (20 years old) coppice layer to abandoned stands of overmature (60 years old) coppice, concluding that the abandoned sites contained a significantly higher species richness. In this study, all saproxylic beetle families were taken into account. The higher species richness could be linked to a higher amount of dead wood. In this study, however, both treatments contained only closed-canopy stands.

Our results are most in line with the findings of Bouget (2005) who compared saproxylic beetle richness in recently created and unharvested windstorm gaps to undisturbed controls in French oak-hornbeam forests. He also found a comparable species richness in both situations, but with striking differences in species composition. Likewise, thermophilous species of the families *Cerambycidae* and *Buprestidae* were more present in the gaps than in the controls. He concluded that unharvested gaps, although not necessarily any richer in species, may help to increase the gamma-diversity at the forest scale. This strong increase in Longhorn beetle and Jewel beetle richness and abundance in disturbance gaps (natural or artificial) was also observed for oak stands by Vodka et al. (2009), and occurs also in other tree species such as Norway spruce (Wermelinger et al. 2002).

Indeed, the combination of a dead wood resource, combined with a diversity of sun-exposure, appear to be the key factors determining the beetle richness. Bouget et al. (2013) compared the species richness in over 150 managed and unmanaged oak and beech stands in France, concluding that

dead wood diversity was the key habitat feature associated with species richness and composition of saproxylic beetle communities. Openness of the stand was equally important for the diversity of both common and rare species in oak, but not for rare species in beech.

Management interventions like coppicing intervene by both removing dead wood resources and generating gap openings. Depending on the amounts and composition of the dead wood resource, and the light conditions before and after the intervention, the combined effect can be positive or detrimental to the saproxylic beetle community, and may also change over time. In a young mature mixed oak hardwood forest in South-Carolina (USA), Ulyshen et al. (2004) created several small gaps, leaving most of the so-called “slash” (i.e., harvest residue) on site. They concluded that wood-dwelling insects initially benefited from the disturbance, but that the disturbed areas became less favorable with time. As the amount of dead wood decreased and the canopy re-closed, the abundance of insects decreased. After 6 years, the gaps showed lower abundances than the younger gaps and the surrounding undisturbed forest. If large and diverse dead wood resources are available and natural canopy disturbances are present, the positive effect of active intervention may apparently disappear even for thermophilous species. In Białowieża, Gutowski (1986) found a lower species richness for the thermophilous *Cerambycidae* in the forest areas managed under a shelterwood regime compared with the unmanaged strict forest reserve.

Further detailed and focused research is needed in active and former coppice-with-standards forests to obtain a better understanding of the saproxylic beetle communities of these woodlands, their requirements and the most suitable management regime for their conservation. In the meanwhile, it is advisable to apply differentiated management alternatives in these oak forests. In Meerdaal forest such an approach is applied to date in the mixed oak stands. It includes non-intervention strict reserves, active coppice-with-standards restoration and regular multifunctional silviculture with conservation of veteran trees and a gradual increase in dead wood. We consider this to be an effective approach to conserve and further develop the saproxylic species richness of the site.

## Conclusions

The results of this exploratory survey of saproxylic beetles in the Meerdaal forest showed that the overall species richness in the forest complex was quite high and comparable to other strict forest reserves in Germany and Luxembourg.

The comparison of different sampling sites also provided indications of the effect of coppice-with-standards as compared to non-intervention, on species richness and

composition of saproxylic beetles. Both management options appear to contain an equally species-rich community, including specific rare and endangered species. Species composition in the coppice-with-standards plot is however apparently different from the non-intervention sites and the beech tree alley plot.

The likely effect of the “management” factor, closely linked to light availability, appeared to be more important than the dominant tree species in the plot. These observations are based on an exploratory survey, and need further investigation and a more rigorous sampling design to draw full conclusions. The results do, however, support these preliminary conclusions and may be considered as a pilot for further specific investigations on the topic.

The observations also suggest that a diversified management approach may be the most suitable strategy to conserve and enhance diverse saproxylic beetle communities in formerly intensively managed semi-natural woodlands. This could include areas of non-intervention and active conservation management aimed at producing open-canopy stands with considerable amounts of sun-exposed deadwood, combined with multifunctional forest management conserving dead wood and veteran trees, as already suggested by Fuller & Warren (1996). At the same time, one should always consider what aspects of biodiversity and natural value are desirable. Too strong a focus on maximizing species richness and specific target species may be detrimental to other species and other aspects of natural value like naturalness, authenticity or replaceability. Finally, as Götmark (2013) states, there is often not one correct management option. Focusing on a single option involves the risk of overlooking management alternatives.

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